# **CCR RULE ASSESSMENT OF CORRECTIVE MEASURES (ACM) REPORT**

# **COAL COMBUSTION BYPRODUCT DISPOSAL FACILITY**

Pleasants Power Station Pleasants County, West Virginia

*Prepared for:* 

### **FirstEnergy**

*800 Cabin Hill Drive Greensburg, PA 15601* 

*Prepared by:* 

**Tetra Tech, Inc.** 

*400 Penn Center Boulevard, Suite 200 Pittsburgh, PA 15235 Phone: (412) 829-3600 Fax: (412) 829-3260* 

Tetra Tech Project No. 212C-SW-00070

### **October 2019**

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# **ACRONYMS/ABBREVIATIONS**





# **1.0 INTRODUCTION**

This Assessment of Corrective Measures (ACM) Report was prepared by Tetra Tech, Inc. (Tetra Tech) on behalf of FirstEnergy Generation (FE) for the Coal Combustion Byproduct Disposal Facility ("CCBDF", "CCR unit", or "Site") at the Pleasants Power Station (hereinafter referred to as the "Station"). The Station is located near the town of Belmont in Pleasants County, West Virginia. This report was developed to comply with pertinent requirements of the United States Environmental Protection Agency (USEPA) Coal Combustion Residuals (CCR) Rule, specifically the Assessment of Corrective Measures requirements per 40 CFR § 257.96.

As discussed further below, CCR Rule groundwater Assessment Monitoring (AM) conducted at the Site identified arsenic concentrations in certain downgradient CCR monitoring wells which were at Statistically Significant Levels (SSLs) that exceeded the Groundwater Protection Standard (GWPS) for arsenic, resulting in the need to conduct an Assessment of Corrective Measures per 40 CFR § 257.96.

# **1.1 PURPOSE**

The purpose of this ACM Report is to provide the following: background on groundwater monitoring findings leading to the ACM; an overview of potential corrective measures which were evaluated; and a comparative evaluation of the corrective measures with regards to the pertinent CCR Rule criteria. In addition, the report specifies the path for meeting Selection of Remedy (SoR) requirements of the CCR Rule (per 40 CFR § 257.97). The assessment of corrective measures has included developing and evaluating new field and laboratory information and data as well as reviewing historical field and laboratory information and data developed by other professional engineers and geologists. In preparing this report, Tetra Tech has exercised its professional judgement in accordance with generally accepted engineering and geologic principles and practices to identify and assess the range of potential corrective measures described herein.

# **1.2 REGULATORY REQUIREMENTS**

#### **Initiating and Completing an Assessment of Corrective Measures**

40 CFR§ 257.96(a) requires that within 90 days of finding that any constituent listed in Appendix IV has been detected at a SSL exceeding the GWPS or immediately upon detection of a release from a CCR unit, the owner or operator must initiate an assessment of corrective measures to prevent further releases, to remediate any releases, and to restore affected areas to original conditions. The assessment of corrective measures must be completed within 90 days, unless the owner or operator demonstrates the need for additional time to complete the assessment of corrective measures due to site-specific conditions or circumstances. The 90-day deadline to complete the assessment of corrective measures may be extended for no longer than 60 days.

#### **Characterizing the Nature and Extent of Release**

Following identification that one or more Appendix IV constituents has been detected at a SSL exceeding the GWPS, the owner or operator of the CCR unit must also:

(1) Characterize the nature and extent of the release (N&E) and any relevant site conditions that may affect the remedy ultimately selected. The characterization must be sufficient to support a complete and accurate assessment of the corrective measures necessary to effectively clean up all releases from the CCR unit pursuant to § 257.96. Characterization of the release includes the following minimum measures:



- (i) Install additional monitoring wells as necessary to define the contaminant plume(s);
- (ii) Collect data on the nature and estimated quantity of material released including specific information on the constituents listed in Appendix IV and the levels at which they are present in the material released;
- (iii) Install at least one additional monitoring well at the facility boundary in the direction of contaminant migration and sample this well in accordance with 40 CFR 257.95(d)(1) ; and
- (iv) Sample all wells in accordance with 40 CFR 257.95(d)(1) to characterize the nature and extent of the release.

The following summarizes the timeline pertaining to compliance at the Site with the above CCR Rule requirements:

- February 13, 2019 (Revised April 5, 2019) Pursuant to 40 CFR 257.95(g) and 257.105(h)(8), FE provided notification in the Operating Record that the 2018 groundwater Assessment Monitoring (AM) program at the Site had identified arsenic, barium, fluoride, lithium, and radium concentrations detected at SSLs above their respective GWPSs established as per 40 CFR 257.95(h). Also, at that time, FE initiated activities to characterize the nature and extent of release. The notification was posted to the publicly accessible website on April 5, 2019.
- April 15, 2019 Pursuant to 40 CFR 257.95(g)(3)(i) and 257.105(h)(9), FE provided notification in the Operating Record that an Assessment of Corrective Measures (ACM) had been initiated for the Site. The notification was posted to the publicly accessible website on May 22, 2019.
- July 15, 2019 Pursuant to 40 CFR 257.96(a), FE provided in the Operating Record a demonstration that, based on hydraulic characteristics of the uppermost aquifer, an additional 60 days was required to complete the ACM.

This document was developed to meet requirements of 40 CFR § 257.96(c), which states the following:

"The assessment under paragraph (a) of this section must include an analysis of the effectiveness of potential corrective measures in meeting all of the requirements and objectives of the remedy as described under § 257.97 addressing at least the following:

- (1) The performance, reliability, ease of implementation, and potential impacts of appropriate potential remedies, including safety impacts, cross-media impacts, and control of exposure to any residual contamination;
- (2) The time required to begin and complete the remedy;
- (3) The institutional requirements, such as state or local permit requirements or other environmental or public health requirements that may substantially affect implementation of the remedy(s)."

### **1.3 SITE BACKGROUND**

CCRs produced at the Station are placed in the facility's captive CCBDF, which is located approximately one mile east-southeast of the Station. The facility consists of both a wet disposal area (impoundment) and dry disposal area (landfill) developed in the McElroy's Run watershed. Taken together, the landfill and impoundment are regulated under West Virginia Department of Environmental Protection (WVDEP) Solid Waste/National Pollutant Discharge Elimination System (NPDES) Water Pollution Control Permit No. WV0079171. A WVDEP groundwater monitoring program for the landfill has been in effect since 1994 and a separate CCR Rule groundwater monitoring program has been in effect since 2017. As per the CCR Rule, the landfill and impoundment are considered two separate, existing CCR units that share a



common boundary (the impoundment dam). As provided by the CCR Rule, a multi-unit groundwater monitoring system has been established for the CCBDF.

As shown on Figure 1-1, the impoundment is situated in the upper portion of the watershed and the landfill is situated in the lower portion of the watershed (adjacent to, and overlying, the impoundment dam). The impoundment is unlined and has been in continuous use since the late 1970s, while the landfill is lined and has been in continuous use since the early 1990s. At the current water level, the surface impoundment area is about 250 acres. The impoundment dam was constructed with a clay-filled cutoff trench at the upstream toe and a clay blanket on the upstream slope for a low permeability barrier. The downstream portion of the dam was constructed using compacted fly ash and periodic layers of bottom ash for blanket drains connected to sloping chimney drains that collect seepage to discharge pipes for monitoring. The downstream face of the dam is covered by the landfill facility which WVDEP considers to be a buttress to the dam. The landfill consists of three primary development stages (I, II, and III in the original permit drawings and now referred to as 1, 2, and 3) which are further subdivided into construction subareas (e.g., Stage 1G, 2A, etc.). At this time, development and disposal operations have only been performed in the Stage 1 and 2 areas while the Stage 3 area remains undeveloped. Up until 2009, all of the landfill subareas were constructed with a compacted clay liner system that included an underlying combined groundwater underdrain/leak detection system and overlying leachate collection system. However, since 2009 (in subareas 1G and 2B), a composite geosynthetic liner system (geosynthetic clay liner and geomembrane) has been utilized that also includes an underlying combined groundwater underdrain/leak detection system and overlying leachate collection system. For all portions of the landfill that overlie the downstream face of the impoundment dam, a bottom ash blanket drain layer has also been utilized under the liner system. Leachate and contact stormwater runoff from the Stage 1 and 2 disposal areas are managed in Sedimentation Pond Nos. 1 and 2, which are lined impoundments located immediately down-valley of the future Stage 3 landfill development area.

Groundwater in the CCBDF area occurs primarily within fractured bedrock and flow is controlled primarily by topography with limited, secondary control by orientation (strike and dip) of the rock units. The fractured bedrock of multiple sandstone units which have been collectively identified as the uppermost aquifer for CCR Rule groundwater monitoring for the combined landfill and impoundment units. Historic and recent groundwater level data indicate groundwater flows north from the topographically higher areas located to the south and southeast of the impoundment. West and northwest of the impoundment dam, topography may be the dominant influence on groundwater flow, as the multiple sandstone units underlying the site are eroded and discontinuous across the valley. Groundwater flow northwest of the dam and under the landfill is in the downstream direction of McElroy's Run (toward the west). Flow in all of the rock units exhibit very little seasonal and temporal fluctuations. A representative set of water level data from the time period of this ACM (July 2019) were used for contouring groundwater elevations and identifying flow patterns at the Site (refer to Figure 1-2). These water levels were similar to historical levels across the Site. As such, separate mapping for other time periods was not necessary for this report. A more detailed discussion of the site's geologic and hydrogeologic characteristics can be found in Section 4.0 of this report.

As detailed in the CCR unit's most recent Annual CCR Groundwater Monitoring and Corrective Action Report ("2018 AGWMCA Report", accessible at http://ccrdocs.firstenergycorp.com/), the certified CCR monitoring well network consists of three upgradient (background) wells (GW-7, -21, and -22), seven downgradient wells to monitor the northern side of the combined CCR units (GW-9, -19, -20, -23, -24, -25, and -26), and three downgradient wells to monitor the western side of the combined CCR units (GW-27, - 28, and -29), as shown on Figure 1-1. It is noted there is also a groundwater monitoring well network at the Site associated with the state solid waste permit, and these wells are also shown on Figure 1-1. As discussed in Section 3.0, some of the state network wells were added to the monitoring program for the



N&E characterization since they were strategically located side-gradient and downgradient of the CCR waste boundary wells and are screened in the same monitored aquifer system.

# **1.4 OVERVIEW OF REPORT CONTENTS**

Section 1.0 of this report provided an overview of the CCR ACM regulatory requirements and background on the CCR unit and CCR groundwater monitoring well network. Section 2.0 summarizes Detection and Assessment Monitoring results as well as the findings of the Appendix III ASD and Appendix IV ASD. Section 3.0 summarizes the Nature and Extent of Release Characterization. Section 4.0 presents the Conceptual Site Model (CSM). Section 5.0 provides the identification and screening of remediation technologies to address arsenic SSLs in groundwater, and Section 6.0 presents the assessment of corrective measures by comparing the candidate technologies to ACM criteria in 40 CFR § 257.96(c). Section 7.0 summarizes the Selection of Remedy (SoR) process. Section 8.0 provides references for documents cited in this report.

## **2.0 GROUNDWATER MONITORING RESULTS**

This section summarizes the findings of the Site's CCR Rule Detection Monitoring (DM) program, the associated Appendix III ASD, and the subsequent AM program and Appendix IV ASD which, taken together, led to the requirement to conduct the ACM. Details on each phase of monitoring and the ASDs can be found in the referenced documents and the pertinent Annual Groundwater Monitoring and Corrective Action Reports.

### **2.1 DETECTION MONITORING & APPENDIX III ALTERNATE SOURCE DEMONSTRATION**

### **2.1.1 Detection Monitoring Results**

FE performed the first DM sampling event in September and October 2017. Following receipt of the validated analytical results, a statistical evaluation of the data was completed in January 2018 and the results indicated that there were statistically significant increases (SSIs) for boron, calcium, chloride, fluoride, pH, sulfate and total dissolved solids (TDS) in one or more well comparisons. The DM sampling, analysis, statistical evaluation, and findings were included in the 2018 CCR Annual Groundwater Monitoring and Corrective Action Report, which is available on the Site's publicly accessible CCR website (http://ccrdocs.firstenergycorp.com/).

### **2.1.2 Alternate Source Demonstration**

Following the identification of SSIs in downgradient Site well samples for Appendix III parameters identified in Section 2.1.1, FE performed an ASD per 40 CFR § 257.94(e)(2). The ASD was performed by Tetra Tech, Inc. (Tetra Tech) to determine whether a source other than the CCR unit caused the SSIs or that the apparent SSIs resulted from errors in sampling, analysis, statistical evaluation, or natural variation in groundwater quality. The ASD scope and findings are presented in the Tetra Tech report entitled, "CCR Appendix III Alternative Source Demonstration Report - 2017 Detection Monitoring, McElroy's Run Coal Combustion Byproduct Disposal Facility, Pleasants Power Station," dated April 16, 2018. The subject report was placed in the facility's operating record in April 2018. The Appendix III ASD concluded that there are potential on-site sources which may have contributed to the SSIs for some constituents; however, it was not possible within the scope of work conducted to definitively confirm these sources resulted in all of the SSIs.



Since the ASD did not conclusively determine that all of the SSI constituents were related to sources or conditions other than the CCR unit, in accordance with 40 CFR 257.95(b), the Station transitioned from Detection Monitoring to Assessment Monitoring (discussed in the following section).

### **2.2 ASSESSMENT MONITORING & APPENDIX IV ALTERNATE SOURCE DEMONSTRATION**

FE performed two rounds of Assessment Monitoring at the Site in May and August 2018 (events AM-1 and AM-2, respectively) in accordance with the facility's CCR groundwater monitoring plan. Following receipt of the validated analytical results, FE performed statistical evaluations of the 2018 AM data to determine whether there were any detected Appendix IV parameters with SSLs above the CCR Unit's established GWPSs. Arsenic, barium, fluoride, lithium, and radium were the only parameters detected at concentrations greater than their respective GWPS, as documented in the facility's Operating Record in February 2019. However, subsequent to the AM-1 and -2 statistical evaluations, groundwater level data collected at the Site necessitated a modified interpretation of current groundwater flow patterns along the northern boundary and an associated revision to the upgradient well comparisons in that area. The revised statistical evaluations determined that arsenic SSLs occurred in more wells than previously indicated but that fluoride was no longer an SSL for the single well (GW-20) in which the SSL was identified. As such, fluoride was no longer identified as an SSL and was not evaluated as part of the Appendix IV ASD nor evaluated in this ACM. Additional detail regarding the revised interpretation of groundwater flow patterns at the site and the associated impacts on statistical evaluations of AM data is provided in the Appendix IV ASD report included as Attachment A.

FE subsequently performed the first of the 2019 AM sampling events (AM-3) in February 2019, and the validated data was statistically evaluated in August 2019. The AM-3 results were consistent with the previous results with respect to having SSLs for arsenic, barium, lithium, and radium (SSL data from sampling events AM-1, -2, and -3 are also provided in the Appendix IV ASD report included as Attachment A). The second 2019 AM sampling event (AM-4) was performed by FE in July 2019, but the receipt and statistical evaluation of the validated data was not completed in time to be included with this ACM report. Those findings will be included as part of the CCR unit's 2019 AGWMCA Report. To date, no other Appendix IV constituents have been detected at SSLs above the their GWPS under the facility's AM program.

Pursuant to 40 CFR § 257.95(g)(3)(ii), Tetra Tech performed an ASD to assess if the Appendix IV SSLs determined for events AM-1, -2, and -3 were attributable to a release from the CCR unit or from a demonstrable alternative source(s). As part of the Appendix IV ASD, a single nature and extent of release characterization sampling event was performed in July 2019 that included wells from the state monitoring program (discussion in Section 3.2 below). The Appendix IV ASD determined that the barium and radium SSLs can be attributed to historical and current oil and gas exploration and production activities that have occurred at the Site; that the source of the lithium SSLs are currently indeterminate but there is a high potential they are also attributable to oil and gas impacts at the Site; and that the arsenic SSLs could not be attributed to sources other than the CCR unit. As such, a transition to N&E characterization and ACM for arsenic per § 257.96 of the CCR Rule commenced as discussed in the following sections.

### **3.0 NATURE AND EXTENT OF RELEASE CHARACTERIZATION**

Pursuant to 40 CFR 257.95(g)(1), FE initiated an N&E of release characterization concurrent with performing the Appendix IV ASD. Following confirmation that the arsenic SSLs were not attributed to sources other than the CCR unit, N&E characterization continued and ACM commenced. This section



summarizes the occurrence and fate and migration characteristics of arsenic, N&E activities conducted as part of the CCR Rule requirements, temporal changes in arsenic concentrations in Site leachate and groundwater as well as the extent of arsenic in Site groundwater as identified by the N&E activities.

### **3.1 NATURE OF ARSENIC**

The following is an overview of arsenic sources, its key geochemical properties, and current regulatory concentration limits for health and environmental protection.

## **3.1.1 Arsenic Sources and Key Geochemical Properties**

Arsenic in groundwater can be derived from various natural and anthropogenic sources including CCRs. It can occur in various forms and its concentration and migration characteristics in groundwater are controlled by the properties of aquifer materials and geochemical conditions (e.g., pH, oxidation-reduction potential, presence of competing anions which may inhibit sorption, etc.). A change in downgradient aquifer properties and geochemical conditions can result in potentially changing the mobility and concentration of arsenic. Therefore, the factors which control arsenic concentrations at a given site can be very complex. The following summarizes the occurrence of arsenic and key geochemical properties which affect its fate and migration characteristics that should be considered in site characterization and remediation strategies:

- Natural sources of arsenic are derived from a wide array of geologic materials, including igneous, metamorphic and sedimentary rocks. Arsenic may subsequently be accumulated during secondary mineral formation in overburden materials and soils. In contrast, anthropogenic sources are typically derived from the land application of arsenical pesticides and herbicides and from disposal of arsenic-bearing wastes generated during processing of ore materials for production of commercial products. (USEPA, October 2007).
- The median concentration of arsenic across all coal types is 7.7 mg/kg. Most arsenic associated with bituminous coal is associated with iron sulfides. While arsenic concentrations in coal ash can be in the range of those measured in background soils, typical arsenic levels in fly ash are higher than the typical levels in soils. (EPRI 2010).
- The most common forms of arsenic in groundwater are their oxy-anions, arsenite [As(III)] and arsenate [As(V)]. Under moderately reducing conditions, arsenite is the predominant species. In oxygenated water, arsenate is the predominant species. Both anions are capable of adsorbing to various subsurface materials, such as ferric oxides and clay particles. Ferric oxides are particularly important to arsenate fate and transport as ferric oxides are abundant in the subsurface and arsenate strongly adsorbs to these surfaces in slightly acidic to neutral waters (USEPA CLU-IN website).
- Arsenic mobility is lowest at pH 3 to 7 and increases at very acidic or alkaline pH (EPRI 2010). At higher alkaline pH, sorption still occurs, but to a lesser degree. Hence, under alkaline conditions, arsenate/arsenite can be expected to be more mobile. The arsenic oxy-anions are also sensitive to redox conditions, and the dominance of arsenate versus arsenite will change with this changing redox. Arsenic can also complex with organic compounds, which can affect its mobility.
- The extent to which inorganic arsenic will partition to mineral surfaces will also be affected by the competition of sorption sites with other anions in solution. There are several commonly occurring anions in natural waters (e.g., phosphate and sulfate) that can compete with arsenic sorption to mineral surfaces. These competitive sorption reactions will be active for all arsenic aqueous species in oxidized and reduced systems.



• Arsenic-bearing colloidal material may be mobilized either from changes in the surface charge on colloids or through deflocculation and suspension of colloidal material through dissolution of cementing agents within the aquifer matrix. Both processes would be facilitated in aquifers impacted by organic contaminants where microbial activity may be stimulated resulting in the generation of reducing conditions and/or the production of low molecular weight organic compounds that partition to fine-grained sediments. (USEPA, October 2007)

### **3.1.2 Regulatory Concentration Limits for Health and Environmental Protection**

Research into state and federal drinking water, National Pollutant Discharge Elimination System (NPDES), and environmental standards by Tetra Tech found the following with respect to concentration limits:

- The federal Maximum Contaminant Level (MCL) for arsenic in drinking water was revised from 0.05 milligrams per liter (mg/L) to 0.01 mg/L, which is the GWPS in effect at the Site.
- For non-potable water sources, federal ambient water quality criteria (AWQC) have been developed that are protective of aquatic life. For arsenic, current statutes list both acute and chronic criteria for arsenic in fresh waters as 0.34 mg/L and 0.15 mg/L, respectively (USEPA, October 2007).
- West Virginia water quality criteria are determined by the state's water use category assigned to the receiving water which, for arsenic, varies from 0.01 mg/L (for public water supply or recreational water contact use) to 0.1 mg/L (for propagation and maintenance of fish and other aquatic life). In those instances where a receiving water does not have a use category assigned, the protective concentration limits for human contact and public water supply (0.01 mg/L) are used. There are also separate criteria for arsenite [As(III)] that apply to aquatic life only and vary between 0.15 mg/L (chronic limit) and 0.34 mg/L (acute limit), which align with the federal AWQC criteria noted above.

### **3.2 NATURE AND EXTENT OF RELEASE CHARACTERIZATION ACTIVITIES**

In an effort to characterize the nature and extent of arsenic in groundwater at the Site and gather information which could be helpful in evaluating potential corrective measures, the following activities were conducted by Tetra Tech in 2019.

# **3.2.1 Additional Monitoring Points**

As previously noted, there are several monitoring wells and piezometers present at the Site that are part of the WVDEP groundwater monitoring system but are not part of the CCR monitoring network (the basis for the CCR monitoring network development is presented in detail in Tetra Tech, 2017). The locations of these wells and piezometers are shown on Figure 1-1 and they either monitor stormwater and/or leachate ponds at the Site (these types of ponds are not required to be monitored by the CCR Rule), the landfill or the impoundment but are positioned too far from the waste boundary to meet the CCR Rule location criteria, or they are currently inactive because they're situated adjacent to the current waste boundary but slated for decommissioning during future permitted expansion of the waste boundary. Referring to Figure 1-1, these wells include GW-3, GW-4, GW-5, GW-8, GW-12, GW-17, MP-1B, MP-3, and MP-4, and the piezometers include P-96-1, -2, -4, and -5. Based on groundwater flow patterns at the Site and proximity to the facility boundary, it was determined that CCR downgradient monitoring wells GW-9, -19, -20, -23, - 24, -25, and -26 fulfilled the requirement of 40 CFR § 257.95(g)(3)(iii) of having at least one monitoring



well positioned at the facility boundary in the direction of contaminant migration (refer to Figure 1-2). As such, both the CCR and/or non-CCR monitoring wells and piezometers were used for N&E of release characterization and no additional monitoring wells have thus far been installed.

### **3.2.2 N&E Sampling and Analysis Program**

As previously noted in Section 2.2, two rounds of regularly scheduled AM sampling (AM-3 and AM-4) were performed in 2019 for the CCR Rule monitoring network with the samples being analyzed for Appendix III parameters and all Appendix IV parameters. As also noted in Section 2.2, as part of the Appendix IV ASD work, a third sampling event, concurrent with the AM-4 event, was performed specifically for the N&E monitoring points described in Section 3.2.1 with the samples analyzed for Appendix III parameters and for arsenic, barium, fluoride, lithium, and radium. Laboratory analysis and data validation activities were completed for the AM-3 sample set in August 2019 but remain in progress for the AM-4 and N&E sampling event data sets. As such, the currently available findings (sampling events AM-1, -2, and -3) are presented in the following section; the AM-4 and N&E results were unable to be incorporated into this ACM, but preliminary review of the data indicates concentration trends similar to previous sampling events. The AM-4 and N&E sampling event findings will be included as part of the CCR unit's 2019 AGWMCA Report. To date, no other Appendix IV constituents have been detected at SSLs above the their GWPS under the facility's AM program.

# **3.3 EXTENT OF ARSENIC AND TRENDS IN CONCENTRATION**

Figure 3-1 presents time series analysis showing total arsenic concentrations detected in groundwater from April 2005 to February 2019. Also shown for reference is a line indicating the 0.01 mg/L arsenic GWPS. As indicated, prior to adding groundwater monitoring wells as part of the CCR Rule compliance work in 2016, the wells with the highest concentrations were MP-1B, GW-3, and GW-4. Since implementation of groundwater monitoring as part of the CCR Rule compliance work in 2016 (including installing new monitoring wells GW-19 through GW-29), GW-19 and GW-22 have typically been the wells having the highest arsenic concentrations. Both of these wells show substantial seasonal fluctuations in arsenic concentrations over the monitoring period.

Figures 3-2 and 3-3 are iso-concentration maps representative of the areal distribution of total arsenic in groundwater in the monitored CCR aquifer for April 2017 and February 2019, respectively. Concentrations greater than the arsenic GWPS of 0.01 mg/L for the aquifer are shaded on the maps. It is noted that while arsenic concentration results are posted for each monitoring well, certain wells (specifically GW-5 and GW-20) which are not screened in the Grafton Sandstone or believed to be hydraulically connected to it, are excluded from contouring of arsenic values (these wells have much higher hydraulic heads than the nearby Grafton Sandstone wells). As discussed below in Section 4.1, the Grafton Sandstone is the monitored aquifer at the site. GW-5 and GW-20 are screened in intervals (Lower Connellsville Sandstone / Lower Clarksburg Redbeds) which are situated above the Grafton Sandstone. The wells were screened in these intervals because they are the shallowest aquifer units adjacent to the CCR unit in these areas. However, it is noted that neither GW-5 or GW-20 had reported concentrations above the GWPS during their May 2017 and April 2019 sampling events, which were close in time to the above-referenced April 2017 and February 2019 sampling events.

Based on interpolation of concentration gradients between the well measurements, both figures show elevated arsenic concentrations occurring through the impoundment and nearby adjacent areas, with the highest concentrations occurring at GW-19 (northwestern area) and GW-22 (southeastern area) for the April 2017 and February 2019 events. It is noted that there are no groundwater monitoring wells available in the central site area (i.e., beneath the impoundment) which precludes confirming the level of arsenic in the monitored aquifer throughout the central portion of the site. Based on the interpreted distribution in



groundwater, arsenic concentrations above the GWPS occur beyond the property boundaries to the north and southeast. In response to these findings, additional N&E of release characterization work is recommended to determine the extent of arsenic concentrations above the GWPS off-site and to gather information to evaluate geochemical conditions to help model potential for natural attenuation to reduce arsenic concentrations in downgradient offsite areas. These and other additional data needs that are part of the final Selection of Remedy at the Site are discussed in Section 7.2 of this report.

## **4.0 CONCEPTUAL SITE MODEL**

# **4.1 HYDROGEOLOGIC CHARACTERISTICS**

This section provides an overview of hydrogeologic characteristics at the Site based on previous studies as well as more recent work completed under the CCR Rule monitoring program. A more detailed discussion of the site's geologic and hydrogeologic characteristics can be found in the "CCR Groundwater Monitoring System Evaluation Report, Harrison Power Station CCB Landfill", Tetra Tech, October 2017.

Groundwater at the Site is derived from precipitation infiltration, however, infiltration through the CCBDF itself is considered to be minimal to none. The entire landfill footprint is underlain with either a compacted clay or composite geosynthetic liner system, and leachate from the landfill is discharged to lined sedimentation ponds. For the disposal impoundment, the upstream face of the dam is clay-lined and keyed into bedrock and water from the impoundment is continuously discharged through an outflow tower and a siphon system. Leakage from the impoundment to groundwater has previously been interpreted to be negligible due primarily to the occurrence of low permeability redbed units present in the former stream valley floor, but sandstone unit outcrops are also present in the valley floor allowing for infiltration into (and/or out of) those units. Leakage from the impoundment may also be limited by the lacustrine deposition of the CCRs and their subsequent compression into a less permeable layer along the former valley bottom and lower sideslopes in the impoundment pool area.

Groundwater in the CCBDF area occurs primarily within the fractured bedrock of the Conemaugh Group, principally in the following sandstone units (in descending order): Morgantown Sandstone, Grafton Sandstone, Jane Lew Sandstone, and the Saltsburg Sandstone. Groundwater has also been identified in the Ames Limestone and Harlem Coal (in association with the Jane Lew sandstone), and, to a lesser extent, the redbed units at the site. Detailed review of occurrence of groundwater in the CCBDF area indicates that the Grafton Sandstone, often in combination with adjacent hydraulically connected stratigraphic units, is the primary aquifer monitored at the site as part of the CCR monitoring network. Groundwater flow at the CCBDF occurs primarily through networks of interconnected fractures formed through tectonic and stress relief processes. Generally, fine-grained rock units (e.g., redbeds) typically serve as aquitards to limit vertical groundwater migration, while coarser grained rock units (e.g., sandstones) typically have more well-developed and open fracture systems and are the primary conduits for groundwater migration. Infiltrated groundwater moves vertically until relatively low-permeability layers are encountered, where a perched water table forms. The perched groundwater flows laterally towards groundwater discharge points within the former stream valleys (manifested as springs or seeps). A portion of the groundwater also migrates through localized, vertically transmissive fractures that penetrate through the low permeability layers to underlying rock units.

Historic and recent groundwater level data indicate groundwater flows north from the topographically higher areas located to the south and southeast of the impoundment. West and northwest of the impoundment dam, topography may be the dominant influence on groundwater flow, as the multiple sandstone units underlying the site are eroded and discontinuous across the former valley. Groundwater flow northwest of the dam and under the landfill is in the downstream direction of McElroy's Run (toward



the west). Flow in all of the rock units exhibit very little seasonal and temporal fluctuations. A representative set of water level data from the time period of this ACM (July 2019) were used for contouring groundwater elevations and identifying flow patterns at the Site (refer to Figure 1-2). These water levels were similar to historical levels across the Site. As such, separate mapping for other time periods was not necessary for this report. A more detailed discussion of the site's geologic and hydrogeologic characteristics can be found in Section 4.0 of this report.

Hydrogeologic properties for the CCBL area have been estimated as part of previous studies (referenced in Tetra Tech, October 2017). Estimates of hydraulic conductivity (K) are available for the landfill waste materials, natural soils, and bedrock. The estimates are based on limited testing data and should be considered generalized estimates only, particularly for the bedrock, as individual fractures in fractured rock groundwater flow systems typically vary widely in water-yielding capabilities. Estimated K values for landfill waste are in the range of 0.03 feet per day, while remolded K values for the natural soils present across the site (mostly silt/clay) range from  $10^{-4}$  to  $10^{-5}$  feet per day. Based on slug tests in well borings, bulk hydraulic conductivities of bedrock range from 0.5 feet per day (Pittsburgh Redbeds) to 255 feet per day (Morgantown and Saltsburg Sandstones). Slug tests measure the overall K of the tested portion of a boring, so it is likely that discrete fracture K values are much higher than the overall average. Historical packer tests and falling head tests yielded hydraulic conductivity values of 0.003 to 0.3 feet per day for the Saltsburg/Buffalo Sandstones.

Appendix B provides a generalized geologic cross-section completed as part of the solid waste permit application for the site. Cross-Section A-A' is a generally northwest-southeast section extending from the Ohio River to the facility boundary (near the location of CCR well GW-22). The section cuts through the landfill, dam, and impoundment areas and depicts the stratigraphic positioning of the Grafton and Saltsburg Sandstones, the Birmingham and Pittsburgh Redbeds, and the Ames Limestone.

# **4.2 POTENTIAL RECEPTORS**

Based on information contained in the CCBL's recent state solid waste permit renewal applications, there are two downgradient water supply wells located within one mile of the landfill perimeter (this includes areas upgradient, side-gradient, and downgradient of the CCR unit). The study area and well locations are shown on attached Figure 4-1. Referring to this figure, the two wells are located approximately 1,500 to 2,000 feet northwest of the facility boundary and are situated close to the Ohio River. Given that there's a mix of arsenic concentrations at the closest downgradient facility boundary wells, with GW-9 being below the GWPS and GW-19 being above the GWPS, there is potential that attenuation of arsenic concentrations may occur over the relatively long flow path from the GW-9 area to the water supply wells. In addition, given the horizontal proximity of the two water supply wells to the Ohio River, it is likely that both wells draw their water from the Ohio River alluvial aquifer. This is a very high-yield aquifer that would significantly dilute any upland groundwater flows that discharge into it.

# **4.3 SUMMARY OF CSM**

Figures 4-2 and 4-3 are generalized cross-sections presenting the Site CSM, with Figure 4-2 representing the portion of flow that branches off to the northwest and Figure 4-3 representing the portion of flow that branches off to the northeast. In summary, the CSM consists of arsenic leaching from the impounded CCRs at the Site and entering groundwater at the base of the former McElroy's Run valley. A significant volume of leachate and infiltration is removed from the groundwater system by the leachate collection and chimney drain systems present in the lined portions of the landfill and under the impoundment dam, respectively. These flows are collected and routed through the lined sedimentation ponds before being discharged off-site. As the remaining impacted groundwater flows downgradient of the CCR unit it is



expected to undergo attenuation based on a combination of advection, dispersion, and, potentially natural dilution resulting in concentrations that are anticipated to be below the arsenic GWPS before flow reaches a potential receptor.

### **5.0 IDENTIFICATION AND SCREENING OF REMEDIATION TECHNOLOGIES**

Technologies for the treatment of arsenic in groundwater are primarily based on ex-situ or in-situ approaches. Pump-and-treat technologies make use of processes common to water and wastewater treatment for removal of dissolved arsenic. In-situ treatment technologies are less common, but there is emerging research based on the application of permeable reactive barriers for arsenic removal from ground water. This technology is based on installation of reactive solid material into the subsurface to intercept and treat the contaminant plume (USEPA, October 2007). Monitored Natural Attenuation (MNA) may also be appropriate at some sites depending on aquifer properties and geochemical conditions. This section identifies the remediation technologies which were evaluated as part of this ACM and summarizes each technology including associated advantages and disadvantages. The technologies include those pertaining to source control and those addressing the impacted groundwater downgradient of the CCBDF.

# **5.1 SOURCE CONTROL**

When remediating impacted groundwater, controlling on-site sources of historical, current, and future contamination to the aquifer are key components to the overall remediation plan. Source control includes a range of potential actions such as treatment in-place, removal, or containment, or some combination of these actions with the goal of reducing or eliminating, to the extent practicable, future releases. For each of the source control technologies below, the focus has been placed on the disposal impoundment as it's an unlined CCR unit. The landfill is a lined CCR unit that includes a leachate collection system and an underlying combined leak detection/groundwater underdrain system and there have been no indications of any releases from the landfill since it was first developed.

### **5.1.1 Treatment in Place**

For an unlined wet disposal impoundment like the existing CCR unit, options for in place source treatment would include amending the CCRs to reduce their permeability and/or chemically fixate the contaminants of concern and prevent them from leaching out. Amendment of the in-place CCRs would be accomplished by the use of drilled pressurized injection wells or deep auger mixing to introduce an amending agent slurry (e.g., Portland cement). Considering the surface area and volume of materials present in a large impoundment like the CCR unit, implementation of such treatment in-place technologies is impractical and has only been noted herein for completeness in presenting options.

### **5.1.2 Removal**

Source removal for a wet disposal impoundment would require excavating, drying/stabilizing, loading and hauling all of the CCRs currently located in unlined areas and placing them in existing or new on-site or off-site lined disposal areas. In general, advantages of removal include:

- Effectively eliminates the potential for future contamination to occur; and
- Can oftentimes reduce the timeframe over which remediation goals can be attained.

In general, disadvantages include:



- An increased overall risk to cleanup workers, the surrounding community, and the environment due to factors such as fugitive dust generation and heavy construction equipment emissions;
- If off-site transport and disposal is required, an increased potential for severe cross-media environmental effects and safety hazards due to accidents; and
- For a large volume site, removal activities could take an unreasonable amount of time to complete and be financially infeasible.

Given the volume of materials present in a large impoundment like the CCR unit and the corresponding effects that the disadvantages noted above would entail for a facility of such size, implementation of CCR removal from unlined areas at the site is impractical and noted herein for completeness in presenting options.

## **5.1.3 Containment**

Source containment approaches for a wet disposal impoundment would include the construction of a final cover (capping) system and/or the installation of a subsurface cutoff wall. Construction of a final cover system atop all exposed CCR surfaces would eliminate source material releases due to stormwater erosion or fugitive dust generation and would reduce leachate generation by minimizing the infiltration of storm water into the underlying CCRs. Installation of a low permeability upgradient groundwater cutoff wall by trench excavation and/or drilled high pressure injection grouting would minimize source contaminant mobilization by preventing groundwater flow into or through the landfilled CCRs.

In general, advantages include:

- Implementation can usually be completed in a relatively short period of time, depending on the dewatering characteristics of the CCRs and the size and depth of the impounded wastes;
- Final cover system design and construction have established processes;
- Can oftentimes reduce the timeframe over which remediation goals can be attained; and
- Effectively reduces the potential for future contamination to occur.

In general, disadvantages include:

- For cutoff walls, subsurface conditions must be favorable across the Site in order to construct an effective and reliable groundwater flow barrier (this is particularly difficult for controlling fractured bedrock flow);
- Depending on the impoundment size and material depths, final cover systems can be difficult to design with respect to tolerating settlement and maintaining reliable long-term stormwater collection and conveyance controls; and
- Final cover systems require routine monitoring, maintenance, and repair throughout their service life.

Given both the large size and the geologic and hydrogeologic characteristics of the Site, the installation of an effective groundwater cutoff wall is impractical and is noted herein for completeness in presenting options. However, construction of a final cover system (either a soil-only or typical regulatory composite cap) is a viable option for the CCR unit and is required under the solid waste permit issued by WVDEP for the Site after the impoundment reaches design capacity and is closed.



# **5.2 GROUNDWATER EXTRACTION AND TREATMENT**

Groundwater extraction and treatment (also referred to as "pump and treat") can be used as a containment strategy at or near the source of contamination or to reduce or eliminate the downgradient migration of a plume. The technology accomplishes a certain amount of mass removal from the plume. In its simplest form, extraction and treatment involves the installation and pumping of vertical extraction wells with the extracted water treated for the contaminant(s) of concern using methods appropriate for the type of contaminant (e.g., air stripping for volatile organic compounds, chemical precipitation for certain inorganic compounds, etc.). As with most remedial technologies it is most effective following source control. In most cases the groundwater treatment results in a need to manage residuals (e.g., sludges, filters, etc.) which may also act as a source of contamination if not properly managed. Extraction and treatment system application often has associated contaminant "rebound" effects related to desorption of additional contaminant mass from aquifer materials following the initial extraction phase. Groundwater extraction and treatment can also be accomplished via horizontal wells.

In general, advantages include:

- Accomplishes some contaminant mass removal; and
- Can help to protect receptors (e.g., drinking water wells) by preventing migration beyond the extraction wells.

In general, disadvantages include:

- Likely to have limited success under heterogenous or low permeability aquifer conditions;
- Often requires long term operation and maintenance and power usage;
- Results in treatment residuals which must subsequently be managed; and
- "Rebound" effects can inhibit the ability to achieve remedial goals.

For arsenic, treatment methods include coagulation (i.e., with ferric chloride or alum) and adsorption on packed bed media (e.g., granular ferric hydroxide or activated alumina). Particularly for aluminum-based coagulants and sorbents, the efficiency of arsenic removal can be dramatically enhanced by pre-oxidation of As(III) to As(V). With greensand filtration, the filter media itself is an oxidant and removal of arsenic, whether it occurs in the groundwater as either As(III) or As(V), is enhanced if the groundwater also contains elevated concentrations of Fe(II).

### **5.3 IN-SITU TECHNOLOGIES**

As opposed to technologies such as groundwater extraction and treatment which involve mechanical systems that must be continually operated, "passive" in-situ technologies operate primarily by using a site's natural characteristics (e.g., groundwater flow direction, aquifer geochemical conditions, etc.) to achieve remedial goals. As discussed in this section, in-situ technologies require a strong understanding of an impacted aquifer's physical and geochemical characteristics, which can be "built upon" to achieve remedial goals through adding appropriate reagents to the subsurface environment to achieve contaminant reduction through processes such as adsorption, precipitation, etc.

### **5.3.1 Permeable Reactive Barriers (PRBs):**

A permeable reactive barrier (PRB) typically involves digging a trench perpendicular to groundwater flow and of sufficient depth to intercept a groundwater plume, then placing a reagent in the trench which will react with the impacted groundwater flowing through it in order to reduce contaminant concentrations, primarily through adsorption or precipitation. A funnel and gate type approach can also be utilized for



PRBs where low permeability walls (the funnel) direct groundwater toward a permeable zone containing the reagent (the gate). Some gates are constructed to be readily accessible to facilitate the replacement of the reagent. The reagent is selected based on the constituent of concern and geochemical conditions of the aquifer (e.g. pH and redox conditions).

Certain contaminants are much more amenable to PRB treatment based on their physical and chemical properties. A commonly used reagent is Zero Valent Iron (ZVI) which can be used to convert certain contaminants to non-toxic or immobile species. ZVI has been shown to be effective in treating many halogenated hydrocarbons as well as removing hexavalent chromium, arsenic, and uranium ("Permeable Reactive Barriers, Permeable Treatment Zones and Application of Zero-Valent Iron", USEPA Clu-In Technologies website.) Both As(III) and As(V) can be removed from water by iron wire or filings in batch systems or columns, and this removal has been attributed to sorption and/or surface precipitation of As onto iron oxides (or rust) produced at the metal surface. However, ZVI has not yet been applied in a permeable reactive barrier system for in situ treatment of arsenic-contaminated groundwater. (SERDP, August 2008).

In general, advantages include:

- Essentially a passive type approach (i.e., no continuous operational oversight needed, maintenance is infrequent, etc.); and
- Can be very effective for certain types of contaminants and under the necessary hydrogeologic conditions.

In general, disadvantages include:

- Not suitable for bedrock aquifers;
- Limited by viable trenching depth;
- Suitable reagents have not been proven for all contaminant types (e.g., arsenic); and
- Reactive agent(s) must be replaced on a scheduled basis.

Application of PRB technology at the Site is not considered viable since the uppermost aquifer system occurs along the downgradient northwestern and northeastern flow paths at depths between approximately 100 and 375 feet and includes a fractured bedrock flow component. In addition, Tetra Tech is not aware of any current applications of PRB technology to remediate arsenic in groundwater at CCR sites. As such, it will not be considered in the evaluation of corrective measures discussion in Section 6.0 but could potentially be revisited should additional information about the viability of using this technology at the Site become available during SoR activities.

### **5.3.2 In-Situ Chemical Stabilization via Injection Wells**

In-situ chemical stabilization involves injection into the subsurface via drilled wells a reagent that will result in the precipitation or adsorption of the constituent of concern, and thereby reduce its concentration in groundwater within and downgradient of the injection area. The type of reagent used will depend on the constituent and geochemical conditions within the aquifer including pH, redox conditions, types of natural clays which may be present, etc. It is critical that the aquifer characteristics, particularly permeability, lend themselves to suitable mixing of the reagent with impacted groundwater. Bench scale testing is typically performed to evaluate viability and, if found to be viable, to support design.

In general, advantages include:

• An overall passive approach with minimal disruption of the Site.

In general, disadvantages include:



- Proven reagents are not available for all CCR constituents;
- Changes in geochemistry or aquifer conditions outside of the injection interval may cause certain reactions to "reverse";
- It can be difficult to achieve the desired mixing of the reagent with impacted groundwater under low permeability and/or heterogenous aquifer conditions (e.g., fractured bedrock); and
- The longevity of the reagents can be difficult to forecast.

### **5.4 MONITORED NATURAL ATTENUATION (MNA)**

The following summary of MNA is based on USEPA Directive 9200.4 – 17P "Use of Monitored Natural Attenuation at Superfund, RCRA Corrective Action, and Underground Storage Tank Sites", April 21, 1999.

The term 'monitored natural attenuation'… refers to the reliance on natural attenuation processes (within the context of a carefully controlled and monitored site cleanup approach) to achieve site-specific remediation objectives within a time frame that is reasonable compared to that offered by other more active methods. The "natural attenuation processes" that are at work in such a remediation approach include a variety of physical, chemical, or biological processes that, under favorable conditions, act without human intervention to reduce the mass, toxicity, mobility, volume, or concentration of contaminants in soil or groundwater. These in-situ processes include biodegradation; dispersion; dilution; sorption; volatilization; radioactive decay; and chemical or biological stabilization, transformation, or destruction of contaminants

The USEPA directive lists the following among the advantages and disadvantages of the MNA approach:

Potential advantages of MNA include:

- As with any in situ process, generation of lesser volume of remediation wastes, reduced potential for cross-media transfer of contaminants commonly associated with ex situ treatment, and reduced risk of human exposure to contaminants, contaminated media, and other hazards, and reduced disturbances to ecological receptors;
- Less intrusion as few surface structures are required;
- Potential for application to all or part of a given site, depending on site conditions and remediation objectives;
- Use in conjunction with, or as a follow-up to, other (active) remedial measures; and
- Potentially lower overall remediation costs than those associated with active remediation.

The potential disadvantages of MNA include:

- Longer time frames may be required to achieve remediation objectives, compared to active remediation measures at a given site;
- Site characterization can often be more complex and costly;
- Long-term performance monitoring will generally be more extensive and for a longer time;
- Institutional controls may be necessary to ensure long term protectiveness;
- Potential exists for continued contamination migration, and/or cross-media transfer of contaminants; and



• Hydrologic and geochemical conditions amenable to natural attenuation may change over time and could result in renewed mobility of previously stabilized contaminants (or naturally occurring metals), adversely impacting remedial effectiveness.

In addition to the above USEPA Directive, a companion Directive was also issued: "Use of Monitored Natural Attenuation for Inorganic Contaminants In Groundwater At Superfund Sites", August 2015, USEPA. The Directive discusses a methodology for considering MNA as a remedial strategy for several inorganic constituents (including arsenic) and expands upon the Tiered Analysis Approach for Developing Multiple Lines of Evidence presented in the original 1999 Directive.

## **6.0 ASSESSMENT OF CORRECTIVE MEASURES**

# **6.1 OBJECTIVE**

The objective of this Assessment of Corrective Measures section is to provide a high-level evaluation of each of the viable remediation technologies presented in Section 5.0 with regards to the criteria identified in 40 CFR § 257.96(c) and previously presented in Section 1.2 of this report. These evaluations are summarized below and in Table 6-1. The criteria evaluated in Sections 6.2 through 6.5 are performancerelated, so each of the technologies has been assigned a subjective rating of "Low", "Medium", or "High" based on how they are anticipated to satisfy each criterion. For the criteria evaluated in Sections 6.6 (time to begin and complete remedy) and 6.7 (institutional requirements), subjective ratings of "Short", "Medium", or "Long" and "Minimal", "Moderate", and "Extensive" have been assigned, respectively. As discussed in Section 5.3.1, the PRB technology was not considered viable due to both the aquifer depth and that the primary aquifer type is fractured bedrock; therefore, it is not included in the evaluations below. A more detailed evaluation of technologies leading to a final selection of remedy will be performed and reported during the Selection of Remedy phase as discussed in Section 7 of this report.

# **6.2 PERFORMANCE**

This section discusses the anticipated performance of each technology relative to its ability to achieve remedial goals in consideration of the CSM. Technologies are ranked as "Low", "Medium", or "High" with regard to their effectiveness in reducing arsenic concentrations in groundwater.

# **6.2.1 Source Control**

#### **Containment using Final Cover System – Medium to High**

As discussed in Section 5.1.3, constructing a final cover system atop all exposed CCR surfaces would minimize the infiltration of storm water into the underlying CCRs which would, in turn, reduce both the groundwater flow rates and the total contaminant loading on the monitored aquifer(s). The magnitude and extent of these reductions depend on the type of final cover system(s) utilized at the Site. As per the CCR unit's current Closure Plan (available on the Site's publicly accessible CCR website http://ccrdocs.firstenergycorp.com/), the existing impoundment area will utilize a soil-only cover system once final closure of the unit is initiated. The soil-only cover system provides a medium level of containment performance while a composite cover system, should the design be revised to utilize one, would provide a high level of containment performance.

# **6.2.2 Groundwater Extraction and Treatment**

**Low**. It is anticipated that the performance of a groundwater extraction and treatment system would be poor due to the anisotropic nature and overall low permeability of the aquifer. It is also noted that the



cross-sectional area through the groundwater flow path downgradient of the landfill is wide. Given that and the fact that groundwater flow at the Site is primarily occurring through bedrock fractures, it is likely many extraction wells would be necessary to ensure that all groundwater flow paths were being captured.

### **6.2.3 In-Situ Technologies**

#### **Chemical Stabilization via Injection Wells – Low**

The anisotropic nature and relatively low permeability of the monitored aquifer would make in-situ treatment by injection wells difficult from the standpoint of achieving adequate contact and reagent mixing with the impacted groundwater.

# **6.2.4 Monitored Natural Attenuation (MNA)**

**Medium to High**. As discussed in Sections 3.3 and 4.2, it's believed that attenuation of the arsenic levels down to the GWPS is occurring near the downgradient facility boundary based on interpolation of the measured concentration gradients. In addition, the nearest water supply users in the downgradient flow path are located approximately 1,500 to 2,000 feet from the facility boundary and are likely drawing from the Ohio River alluvial aquifer. Taken together, the anticipated ongoing performance of MNA would be medium when combined with the eventual installation of a soil-only final cover system, but high if it is combined with the eventual installation of a composite final cover system.

# **6.3 RELIABILITY**

Reliability is the anticipated consistency of a technology to function as designed/expected under variable site-specific conditions. Factors which affect reliability can include aquifer variability (e.g., groundwater geochemistry and flow changes) and equipment performance (e.g., power outages and frequency of maintenance activities). Technologies are ranked as "Low", "Medium", or "High" with regard to their effectiveness in consistently reducing arsenic concentrations in groundwater.

### **6.3.1 Source Control**

#### **Containment Using Final Cover System - High**

The soil-only cover system that is proposed for use during final closure will be designed and constructed in accordance with well-established practices. The design could also be modified to use a composite final cover system that incorporates a geomembrane and an upper layer of vegetated cover soil that's comparable to the soil-only cover system. Both systems are expected to be highly reliable as long as they are properly monitored and maintained, which FE will do for the remainder of the landfill's operating life and for the duration of the landfill's post-closure period as required by the state Solid Waste Permit.

### **6.3.2 Groundwater Extraction and Treatment**

**Medium to High**. Extraction and treatment would require proper operation and maintenance (O&M) of extraction well (e.g., pumps) and treatment system equipment in order to maintain reliability. The aquifer system would also need to be evaluated for the presence of high iron and manganese concentrations as these constituents require measures to be taken to prevent fouling and deterioration of pumps and treatment equipment as well as any connecting piping.



## **6.3.3 In-Situ Technologies**

#### **Chemical Stabilization via Injection Wells – Low to Medium**

It is anticipated that since in-situ chemical stabilization of arsenic in a low yield, fractured bedrock aquifer system via injection wells does not seem to be proven, that reliability would be questionable. Beyond concept reliability, the injection system itself would require proper O&M of the well equipment (e.g., pumps) and the surface batching and feed systems in order to maintain operational reliability.

### **6.3.4 Monitored Natural Attenuation**

**Medium to High**. Based on the factors previously discussed in Section 6.2.4, it is anticipated that reductions in arsenic concentrations would be reliable going forward provided it is combined with the eventual installation of either a soil-only or a composite final cover system and confirmation of geochemical conditions which may affect attenuation.

### **6.4 EASE OF IMPLEMENTATION**

Ease of implementation relates to how challenging the technology installation will be considering sitespecific conditions (e.g., degree of aquifer heterogeneity), the complexity of the design effort (e.g., modeling, bench scale and pilot testing, etc.), and the availability of suitable equipment. Technologies are ranked as "Low", "Medium", or "High" with regard to their ease in being installed to begin reducing arsenic concentrations in groundwater.

### **6.4.1 Source Control**

#### **Containment using Final Cover System – Medium to Low**

Either the currently proposed soil-only cover system or a composite cover system would require the development of construction-level drawings and specifications and then have to proceed through the Station's procurement process before construction could commence. Construction would first require dewatering (and possibly treatment) of all free liquids and sufficient pore water to stabilize the impounded CCRs so they could be graded to receive the cover system and to provide positive drainage. Construction of the cover system would then entail the use of commonly accepted materials but nonstandard means and methods due to the physical nature and engineering characteristics of partially and completely saturated CCRs. The ease of completion would also depend heavily on the size of the area(s) being covered and seasonal weather constraints. Because of these factors, ease of installation for either final cover system is considered medium to low.

### **6.4.2 Groundwater Extraction and Treatment**

**Low**. Based on the anisotropic and low permeability nature of the monitored aquifer, it is likely that many groundwater extraction wells would be needed to attempt to capture impacted groundwater. Given both the topography and the number of below and above ground oil and gas conveyance lines in the targeted intercept areas and the interferences they would present, siting the wells in the desired locations would prove extremely difficult. Bench scale testing would also need to be conducted to identify the best reagent(s) for use in removing the arsenic from solution. Such a bench scale testing program would be expected to go through multiple iterations before establishing the treatment program needs. Because of these factors, ease of installation for this system is considered low.

### **6.4.3 In-Situ Technologies**

**Chemical Stabilization via Injection Wells – Low** 



Implementation would likely be very challenging due to identifying the appropriate reagent(s) and "dosing" strategy to effectively and efficiently treat the aquifer due to the anisotropic conditions. It is likely that various phases of bench scale and field pilot testing would be necessary to support the design.

### **6.4.4 Monitored Natural Attenuation**

**Medium to High**. No additional equipment would be necessary for a natural attenuation remedy. There would likely be a need to add a limited number of properly constructed monitoring wells in the downgradient areas along the northern facility boundary to evaluate the program's performance, and this could present significant difficulties due to the topography of this area and the potential need to negotiate monitoring well easements with downgradient property owners.

### **6.5 POTENTIAL IMPACTS OF APPROPRIATE REMEDIES (SAFETY, CROSS-MEDIA AND CONTROL OF EXPOSURE)**

Potential impacts of technologies were evaluated considering the following:

- Safety: The likelihood that illness, injury, or death directly related to the technology would occur during construction or operations. In general, "active" technologies and those requiring significant construction effort were considered higher risk than "passive" technologies and those not requiring significant construction effort.
- Cross-Media: The likelihood that the technology will result in a transfer of contaminants to the air, surface water, or soil, either from a direct discharge or from management of treatment residuals.
- Control of Exposure: The likelihood that that the technology will result in exposure of contaminants to human or environmental receptors either from a direct discharge or from management of treatment residuals.

Technologies are ranked as "Low", "Medium", or "High" with regard to how likely they are to have negative effects for Safety and Cross-Media, and with regard to how well they avoid negative effects for Control of Exposure.

### **6.5.1 Source Control**

#### **Containment using Final Cover System**

Safety Impacts: **Medium to High**. Construction of either a soil-only or composite final cover system would involve both typical and atypical construction risks, both on-site and off-site. Typical risks would include material deliveries and heavy equipment operations, while atypical risks would include excessive settlement and low shear strengths, both of which are commonly associated with dewatered impoundment CCRs. However, after construction is completed, the final cover system would present little to no implementation-related safety impacts.

Cross-Media Impacts: **Low**. Construction of either a soil-only or a composite final cover system atop all exposed CCR surfaces would eliminate source material releases and potential cross-media impacts to the air, ground surface, or surface water due to stormwater erosion or fugitive dust generation.

Control of Exposure: **High**. Construction of either a soil-only or a composite final cover system atop all exposed CCR surfaces would eliminate direct and indirect exposure to the disposed CCRs.

### **6.5.2 Groundwater Extraction and Treatment**

Safety Impacts: **Medium**. Safety risks associated with drilling extraction wells and construction of a treatment facility would exist but could be minimized through implementation of an appropriate health and safety plan. Likewise, some safety risks would be associated with the operation of the treatment system;



however, such risks could be minimized through proper O&M procedures and through implementation of an appropriate health and safety plan.

Cross-Media Impacts: **Medium**. Treatment residuals would need to be managed. In addition, the potential exists for releases from well connections, valves, system piping, and tanks that could impact site soils and potentially groundwater and surface water.

Control of Exposure: **Medium**. Treatment residuals would need to be properly managed to minimize exposure. In addition, the potential exists for exposure to workers and other on-site personnel from any releases which may occur at the well heads, piping, and any storage tanks that are part of the extraction and treatment system.

### **6.5.3 In-Situ Technologies**

#### **Chemical Stabilization via Injection Wells**

Safety Impacts: **Medium** – There would be safety risks associated with drilling injection wells and handling reagent.

Cross-Media Impacts: **Low to Medium** – Would need to confirm that selected reagent would not have negative impacts associated with downgradient groundwater discharge to surface water.

Control of Exposure: **Medium to High** – Will require proper handling procedures for the selected reagent.

### **6.5.4 Monitored Natural Attenuation**

Safety Impacts: **Medium** - Some additional construction or well installation would be necessary under the MNA remedy; there would be safety risks associated with possibly installing a limited number of properly constructed monitoring wells in the downgradient areas along the northern facility boundary to evaluate the program's performance, but this would not present significant safety impacts.

Cross-Media Impacts: **Low to Medium** – As noted in Section 4.3, the Site CSM indicates groundwater from the monitored aquifer flows to the northwest and the northeast. The nearest drainage feature to the northwest appears to be the Ohio River, which is located approximately ½-mile from the facility boundary. Based on a review of aerial imagery, there does not appear to be a downgradient drainage feature that would intercept the Grafton sandstone within one mile of the facility boundary. However, for flow in both directions, it's believed that attenuation of the arsenic levels down to the GWPS is occurring near the northwestern facility boundary based on interpolation of the measured concentration gradients. In addition, the arsenic levels measured in the Site wells are either below or near the state and federal aquatic water quality criteria presented in Section 3.1.2, which would apply to the Ohio River.

Control of Exposure: **High** - No contamination residuals will be generated. As stated in Section 4.2, the closest downgradient water supply users are located approximately 1,500 to 2,000 feet from the facility boundary.

# **6.6 TIME REQUIRED TO BEGIN AND COMPLETE REMEDY**

The anticipated time required to begin and compete a remedy considers factors such as the complexity of the design, construction, and permitting efforts, as well as forecasting how efficient the technology is expected to be in achieving remedial goals in a timely manner. Technologies are ranked as "Short", "Medium", or "Long" with regard to their anticipated time to reduce arsenic concentrations in groundwater.

### **6.6.1 Source Control**

**Containment using Final Cover System** 



Time to Begin Remedy: **Medium. I**t is anticipated that preparation of engineering and construction drawings and documents and contractor procurement would take approximately two years.

Time to Complete Remedy: **Medium to Long.** As previously noted, construction would first require dewatering operations which would then be followed by installation of the final cover system. All of this work would need to be performed using a phased construction approach that would include seasonal (winter) shutdowns, with the total time to complete construction being approximately five to ten years.

### **6.6.2 Groundwater Extraction and Treatment**

Time to Begin Remedy: **Medium.** It is anticipated that one to two years would be required to initiate a groundwater extraction and treatment remedy in order to allow time for modeling to select well locations; to complete well, pipeline and treatment system design and permitting, and to construct the extraction and treatment systems (medium).

Time to Complete Remedy: **Currently Unknown**. Extraction and treatment, while effective at containment in some settings, is often not successful in achieving remedial goals due to "rebound" effects and other field variables that become more defined during system startup and operation.

## **6.6.3 In-Situ Technologies**

#### **Chemical Stabilization via Injection Wells**

Time to Begin Remedy: **Medium.** Two to three years are estimated for bench scale testing in order to select the treatment reagent(s), perform modeling to identify injection well locations, complete well and injection system design and permitting, and to install the injection wells and construct the injection system (medium).

Time to Complete Remedy: **Currently Unknown**. The time required to complete the remedy will depend on the duration of leaching of arsenic into the aquifer, which is expected to decrease as the CCR unit is covered/capped. The duration of treatment required is difficult to estimate until at least bench scale testing is performed on the selected reagent.

### **6.6.4 Monitored Natural Attenuation**

Time to Begin Remedy: **Short**. As previously noted, it's believed that attenuation of the arsenic levels down to the GWPS is occurring near the northwestern facility boundary based on interpolation of the measured concentration gradients.

Time to Complete Remedy: **Long**. Additional monitoring and the installation of additional monitoring well locations would be necessary to confirm that the GWPS is being attained near the facility boundary. Ongoing monitoring to confirm the remedy continues to be effective would also be proposed with the duration to be determined as part of the Selection of Remedy process discussed in Section 7.0 of this report.

### **6.7 INSTITUTIONAL REQUIREMENTS (STATE AND LOCAL PERMITS AND OTHER APPROVALS)**

Institutional requirements pertain to the anticipated state and local permits and other approvals needed to construct and operate the remedial technology. These can include programs already in-place for a given CCR unit (e.g., solid waste permit) that will need to be modified to accommodate a potential technology, or new programs that may result from a potential technology (e.g., NPDES permit). FE will continue to provide CCR Rule program notifications to WVDEP as required by 40 CFR § 257.106 and will also consult with WVDEP to confirm anticipated permitting requirements that would be associated with the



selected remedy. As mentioned in Section 1.3, the CCBDF is permitted under the WVDEP solid waste regulations; therefore, consultation with the agency will be required to support remedy selection, design, and implementation. The following summarizes the expected permits/approvals which may be required by WVDEP or local authorities for each technology and associated rankings of "Minimal", "Moderate", and "Extensive" with regard to the anticipated level of effort that will be needed to obtain them.

### **6.7.1 Source Control**

#### **Containment using Final Cover System – Minimal to Moderate**

Both the existing soil-only cover system and a potential composite final cover system would be regulated under the state-issued Solid Waste Permit. The use of the soil-only cover system in its current operating capacity would only require minimal modifications to the Solid Waste Permit, while the use of a composite cover system would require moderate modifications to the Solid Waste permit.

### **6.7.2 Groundwater Extraction and Treatment**

It is anticipated that either an amendment to the facility's combined Solid Waste/NPDES permit or a new individual NPDES permit will be required for construction and operation of a treatment system. This would likely constitute a moderate to extensive effort. Well locations, piping, and any excavation related to the treatment system would also need to undergo utility clearances.

### **6.7.3 In-Situ Technologies**

#### **Chemical Stabilization via Injection Wells - Moderate**

It is anticipated that only an amendment to the facility's Solid Waste Permit would be required for construction and operation of an injection system.

### **6.7.4 Monitored Natural Attenuation**

No new or amended permits and/or approvals are anticipated from state or local agencies and authorities for an MNA remedy. The implementation of an MNA remedy would only require the regular renewal of the Solid Waste Permit, which would likely constitute a minimal effort.

### **6.8 COMPARATIVE ANALYSIS OF CORRECTIVE MEASURES ALTERNATIVES**

Based on the evaluation of viable remediation technologies presented in Sections 6.1 through 6.7, MNA, combined with source control by the eventual installation of a final cover system, ranks highest among the evaluated options. It ranks medium to high in performance, reliability, ease of implementation, potential safety impacts and potential for residual contamination impacts. Also, additional monitoring of the groundwater network should be conducted to confirm that there are not trend changes that could impact effectiveness. These and other additional data needs that are part of the final Selection of Remedy at the Site are discussed in Section 7.2. It is also noted that it is anticipated that the installation of a final cover system should accelerate the effectiveness of whichever associated corrective measure is selected.



# **7.0 PROCESS FOR SELECTION OF REMEDY**

## **7.1 SELECTION CRITERIA AND SCHEDULE**

As required by 40 CFR § 257.97(a), FE will, as soon as feasible after completion of this ACM, select a remedy that, at a minimum, meets the performance standards listed in 40 CFR 257.97(b) and the evaluation factors listed in 40 CFR 257.97(c). As required by 40 CFR § 257.97(d), FE will specify as part of the selected remedy a schedule(s) for implementing and completing remedial activities. The schedule will require the completion of remedial activities within a reasonable period of time taking into consideration the factors set forth in 40 CFR § 257.97(d)(1) through (d)(6),

## **7.2 ADDITIONAL DATA NEEDS**

In order to select a remedy that is both effective and implementable, additional data collection and analyses will be required as summarized below:

- Installation of additional monitoring wells downgradient of the northwestern and northeastern flow paths to confirm attenuation of arsenic is occurring near the facility boundary, gather geochemical information pertinent to evaluating arsenic natural attenuation, and to monitor the continued effectiveness of the attenuation mechanisms.
- Modeling of the monitored aquifer to further evaluate the MNA alternative to assist in forecasting likely long-term effectiveness and to estimate timeframes for completing remedial activities.
- Additional research into potential reagents for chemical stabilization of arsenic via injection wells as presented in Section 5.3.2.

### **7.3 REMEDY SELECTION PROGRESS REPORTING**

As required by 40 CFR § 257.97(a), FE will prepare a semi-annual report describing the progress in selecting and designing the remedy. One of the semi-annual reports will be included in the forthcoming 2019 Annual Groundwater Monitoring and Corrective Action Report, which will be completed in January 2020.

### **7.4 PUBLIC MEETING**

As required by 40 CFR § 257.96(e), FE will discuss the results of the corrective measures assessment at least 30 days prior to the selection of remedy, in a public meeting with interested and affected parties.

### **7.5 FINAL REMEDY SELECTION**

Upon selection of a remedy, FE will prepare a final report describing the selected remedy and how it meets the standards outlined in Section 7.1. The final report will include a certification from a qualified professional engineer that the remedy selected meets the requirements of the selection criteria and the final report will be placed in the Station's operating record as required by  $\S$  257.105(h)(12).



### **8.0 REFERENCES**

EPRI, 2010. Arsenic in Coal Combustion Products. Technical Brief No. 1021212. December 2010.

- Tetra Tech, 2017. CCR Rule Groundwater Monitoring System Evaluation Report, Pleasants Power Station, Coal Combustion Byproduct Disposal Facility. October 2017.
- Tetra Tech, 2018. 2017 Annual CCR Groundwater Monitoring and Corrective Action Report, Coal Combustion Byproduct Disposal Facility, Pleasants Power Station. January 2018.
- Tetra Tech, 2019. 2018 Annual CCR Groundwater Monitoring and Corrective Action Report, Coal Combustion Byproduct Disposal Facility, Pleasants Power Station. January 2019.
- USEPA, 1999. Use of Monitored Natural Attenuation at Superfund, RCRA Corrective Action, and Underground Storage Tank Sites. April 21, 1999.
- USEPA, 2015. Disposal of Coal Combustion Residuals from Electric Utilities; Final Rule. Federal Register Vol. 80, No. 74, Part II 40 CFR Parts 257 and 261 Hazardous and Solid Waste Management System. April 17, 2015.
- USEPA, 2015. Use Of Monitored Natural Attenuation For Inorganic Contaminants In Groundwater At Superfund Sites. August 2015.



# **TABLES**









Notes:

1. Subjective ratings of "Low", "Medium", or "High" assigned based on how the potential corrective measures are anticipated to satisfy each evaluation criterion:

Performance: Effectiveness in reducing arsenic concentrations in groundwater.

Reliability: Effectiveness in consistently reducing arsenic concentrations in groundwater.

Ease of Implementation: Ease in being installed to begin reducing arsenic concentrations in groundwater.

Safty Impacts: Likelihood that illness, injury, or death directly related to the potential corrective measure would occur during construction or operations.

Cross-Media Impacts: Likelihood that the potential corrective measure will result in a transfer of contaminants to the air, surface water, or soil, either from a direct discharge or from management of treatment residuals.

Control of Exposure: Likelihood that that the potential corrective measure will result in exposure of contaminants to human or environmental receptors either from a direct discharge or from management of treatment residual

2. Subjective ratings of "Short", "Medium", or "Long" assigned with regard to the anticipated time for each potential corrective measure to reduce arsenic concentrations in groundwater, accounting for factors such as the c

3. Subjective ratings of "Minimal", "Moderate", and "Extensive" assigned with regard to the anticipated level of effort that will be needed to obtain the permits/approvals which may be required by WVDEP or local authoritie

# **FIGURES**





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# **APPENDIX A**

**Appendix IV Alternative Source Demonstration Report – 2018/2019 Assessment Monitoring** 



# **CCR Rule Appendix IV Alternative Source Demonstration Report – 2018/2019 Assessment Monitoring**

# **McElroy's Run Coal Combustion Byproduct Disposal Facility**

Pleasants Power Station Pleasants County, West Virginia

*Prepared for:* 

### **FirstEnergy**

*800 Cabin Hill Drive Greensburg, PA 15601* 

*Prepared by:* 

**Tetra Tech, Inc.** 

*400 Penn Center Boulevard, Suite 200 Pittsburgh, PA 15235 Phone: (412) 829-3600 Fax: (412) 829-3260* 

Tetra Tech Project No. 212C-SW-00070

### **October 2019**

## **CCR RULE APPENDIX IV ALTERNATIVE SOURCE DEMONSTRATION REPORT 2018/2019 ASSESSMENT MONITORING**

**McELROY'S RUN COAL COMBUSTION BYPRODUCT DISPOSAL FACILITY** 

## **PLEASANTS POWER STATION PLEASANTS COUNTY, WEST VIRGINIA**

**Prepared for:** 

**FirstEnergy** 

**800 Cabin Hill Drive Greensburg, PA 15601** 

**Prepared by:** 

**Tetra Tech, Inc. 400 Penn Center Boulevard, Suite 200 Pittsburgh, PA 15235 Phone: (412) 829-3600 Fax: (412) 829-3260** 

**Tetra Tech Project No. 212C-SW-00070** 

**October 2019** 

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### **1.0 INTRODUCTION/BACKGROUND**

FirstEnergy (FE) owns and operates the coal-fired Pleasants Power Station (hereinafter referred to as the "Station") located in Pleasants County, West Virginia. Coal Combustion Residuals (CCRs) produced at the Station are placed in the facility's Coal Combustion Byproduct Disposal Facility (CCBDF or "CCR unit"), which is located approximately one mile east-southeast of the Station (see Figure 1). The facility consists of both a wet disposal area (impoundment) and dry disposal area (landfill) developed in the McElroy's Run watershed. Taken together, the landfill and impoundment are regulated under West Virginia Department of Environmental Protection (WVDEP) Solid Waste/National Pollutant Discharge Elimination System (NPDES) Water Pollution Control Permit No. WV0079171, and the United States Environmental Protection Agency (USEPA) Disposal of Coal Combustion Residuals from Electric Utilities rule (40 CFR Part 257, hereinafter referred to as the "CCR Rule" or "Rule"). As per the CCR Rule, the landfill and impoundment are considered two separate, existing CCR units that share a common boundary (the impoundment dam). As provided by the CCR Rule, a multiunit groundwater monitoring system has been established for the CCBDF.

In accordance with § 257.94 of the Rule, the initial Detection Monitoring (DM) sampling and analysis event for the CCR unit was completed in October 2017, and the statistical evaluation of the resulting data was completed in January 2018. As required by § 257.90(e), the results and findings from the 2017 groundwater monitoring program were documented in the 2017 Annual Groundwater Monitoring and Corrective Action Report (AGWMCA Report) that was posted in both the CCR unit's operating record and on its publicly accessible website in January 2018 (Tetra Tech, 2018). In that report, Statistically Significant Increases (SSIs) for boron, calcium, chloride, fluoride, pH, sulfate, and total dissolved solids (TDS) were determined in several downgradient monitoring wells. Based on the various parameters for which SSIs were identified, an Appendix III Alternative Source Demonstration (ASD) was undertaken as discussed in the 2018 AGWMCA Report (Tetra Tech, 2019). However, all of the Appendix III SSIs that were identified for DM-1 could not be attributed to alternative sources.

During the transition period between completing the statistical evaluation of the DM-1 data and performing the Appendix III ASD, FE performed another round of DM sampling (event DM-2) in order to have data available should the ASD prove to be successful and the facility remained in the DM program. DM-2 sampling occurred in February 2018, with laboratory analysis and data validation completed by April 2018. However, before statistical evaluation of the DM-2 data



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commenced, it was determined that a transition to Assessment Monitoring (AM) was required which precluded the need to statistically evaluate the DM-2 data. As such, a transition to the applicable requirements of AM per § 257.95 of the CCR Rule commenced.

In accordance with 40 CFR § 257.95(b) and (d)(1), two AM sampling events (AM-1 and AM-2) were performed in May and August 2018. Pursuant to §§ 257.94(e)(3), 257.105(h)(5), and 257.106(h)(4), a notice was posted to the facility's Operating Record and issued to the WVDEP in August 2018, to provide notification that a groundwater Assessment Monitoring program for the CCR unit had been established. Pursuant to  $\S$  257.107(h)(4), the subject notice was posted to the facility's publicly accessible website in September 2018. Analytical data summary tables and a description of the 2018 AM program results can be found in the 2018 AGWMCA Report (Tetra Tech, 2019). Once initiated, the AM program continued in 2019 with two additional sampling events performed in February (AM-3) and July (AM-4).

Statistical evaluation of the AM sampling events was completed in January 2019 for AM-1 and - 2 and in August 2019 for AM-3 (validated AM-4 results were not available in time to be included in this report). The statistical evaluations indicated Appendix IV constituent concentrations in downgradient wells at Statistically Significant Levels (SSLs) above applicable Groundwater Protection Standards (GWPS). In accordance with 40 CFR § 257.106(h)(6), a notice was prepared and posted to the facility's Operating Record, issued to the WVDEP, and then posted on the facility's publicly accessible website in April 2019, to provide notification of the SSLs for arsenic, barium, fluoride, lithium, and radium at the CCR unit.

During this same notification period and in accordance with 40 CFR § 257.95(g)(3)(ii), an Appendix IV ASD was initiated to assess if the SSLs determined for the AM-1, AM-2, and AM-3 events were attributable to a release from the CCR unit, from a demonstrable alternative source(s), or if they resulted from errors in sampling, analysis, statistical evaluation, or natural variation in groundwater quality. Pursuant to  $\S$  257.95(g)(4), if a successful ASD has not been completed within 90 days from the date of determining that an SSL has occurred, the CCR unit owner or operator must initiate an Assessment of Corrective Measures (ACM) in accordance with 40 CFR § 257.96. Due to the additional monitoring points, sampling events, laboratory analyses, and evaluations needed to complete a successful ASD, the work to complete the ASD had to be extended. Therefore, and in accordance with 40 CFR  $\S$  257.106(h)(7), a separate notice was prepared and posted to the facility's Operating Record, issued to the WVDEP, and then posted on the facility's publicly accessible website in April 2019, to provide notification of the initiation of



the assessment of corrective measures for arsenic, barium, fluoride, lithium, and radium at the Site.

Subsequent to the above-referenced AM notifications, additional rounds of groundwater level data were collected and evaluated which resulted in a modified interpretation of current groundwater flow patterns along the northern boundary of the Site than were described in the *CCR Rule Groundwater Monitoring System Evaluation Report for the Pleasants Power Station* (Tetra Tech, 2017). In the subject report there were two, separate upgradient/background wells identified for the western and northern boundaries of the CCR unit. The current understanding of groundwater flow based on the additional rounds of groundwater level measurements is such that one upgradient well, GW-7, is now considered the upgradient/background well for both the western and northern boundaries of the CCR unit (Figure 2). This change in groundwater flow pattern is likely attributable to the low permeability of the formation and long stabilization period required for the wells installed along the northern boundary. As such, the AM statistical evaluations that have recently been conducted have incorporated upper prediction limits (UPLs) associated with GW-7 for both boundaries.

The table shown on the following page summarizes the results of the statistical evaluation of the CCR Rule Appendix IV parameters based upon utilizing the updated groundwater flow interpretation (i.e., utilizing the GW-7 UPL for comparison with downgradient constituent concentrations) and lists which wells (labeled "GW-#") have parameters that were determined to be above their GWPS. The revised statistical evaluation based on the updated understanding of groundwater flow patterns determined that arsenic SSLs occurred in more wells than previously indicated (due to the lower arsenic GWPS for MW-7), but that fluoride was no longer an SSL in the single well it was previously found in (GW-20) due to the higher fluoride GWPS for MW-7. As such, fluoride is no longer considered an SSL and was not evaluated in this ASD. A detailed discussion of the revised interpretation of groundwater flow patterns at the site and the associated impacts on statistical evaluations of AM data will be provided in the forthcoming 2019 AGMCA Report that will be issued in January 2020.

After initiating the ACM in April 2019, the ongoing ASD activities were continued as they indicated a strong possibility that the barium, lithium, and radium SSLs were attributable to demonstrable alternative source(s). As such, this ASD report has been prepared to document the evaluation of the AM-1, -2, and -3 Appendix IV SSLs and to incorporate the findings into the CCR unit's ACM.





Note: Downgradient well GW-26 was not sampled (n/s) during the AM-1 and AM-2 events due to insufficient available water.



### **2.0 APPROACH**

For this ASD, a multiple Line of Evidence (LOE) approach as presented in *Guidance for Development of Alternative Source Demonstrations at Coal Combustion Residual Sites* (EPRI, 2017) was followed. This approach divides LOEs into five separate ASD categories (types):

- Sampling causes (ASD Type I);
- Laboratory causes (ASD Type II);
- Statistical evaluation causes (ASD Type III);
- Natural variation not accounted for in the basic DM statistics (ASD Type IV); and
- Potential natural or anthropogenic sources (ASD Type V).

EPRI (2017) includes detailed checklists that provide a standardized, incremental approach that is followed to determine whether additional LOE evaluations are warranted or not. These checklists include:

- Checklist 1: Sampling, Laboratory, or Statistical Causes (ASD Types I, II, and III);
- Checklist 2: LOEs Associated with the CCR Unit (ASD Type IV); and
- Checklist 3: LOEs Associated with Alternative Natural or Anthropogenic Sources (ASD Type V).

For this ASD all three Checklists were completed and are attached as Tables 1, 2, and 3. Based on indications from these checklists as well as the CCR unit's topographic and geologic setting, development and operational history, and currently available information and data, it was determined that additional evaluations of the following site-specific LOEs were warranted:

- Regional groundwater chemistry studies/reports; and
- Potential existing and historic oil and/or gas production well effects.

The findings from the checklist completion activities and site-specific LOE evaluations are summarized in Section 3.0.



## **3.0 SUMMARY OF FINDINGS**

## **3.1 ASD CHECKLIST 1**

ASD Checklist 1 is attached as Table 1 of this report. The checklist evaluations were performed by re-reviewing the CCR groundwater monitoring program's field sampling notes and chain-ofcustody forms, laboratory data validation (Level 2) reports, statistical evaluation spreadsheets, and results from field-filtered duplicate samples that were obtained during events where turbid unfiltered samples had been obtained. As indicated in Table 1, for many potential sampling, laboratory, or statistical evaluation causes, no instances/issues/indications were identified. Sample contamination with petroleum and/or brine from on-site oil and gas exploration and production activities could be a contributing factor for the SSIs and SSLs for barium, lithium, and radium in GW-23, -24, and -25 (as discussed in Section 3.5 of this report, barium, lithium, and radium have been documented as being associated with oil and gas well brines). For other potential causes where some issues were identified, it was determined that they most likely did not contribute to the Appendix IV SSLs.

Based on these LOE findings, laboratory analysis and statistical evaluations are not demonstrable alternative sources of all the Appendix IV SSLs determined for the AM-1, -2, and -3 events, while sample turbidity and contamination are potential sources of the SSIs and SSLs determined for barium, lithium, and radium in some of the downgradient monitoring wells.

## **3.2 ASD CHECKLIST 2**

ASD Checklist 2 is attached as Table 2 of this report. The checklist evaluations were performed by re-reviewing the groundwater analytical results (background, DM, and AM) for both Appendix III and IV parameters, leachate data for the CCR unit (specifically for arsenic, barium, lithium, and radium) provided by FE, and hydrogeologic and design information and data included in *CCR Rule Groundwater Monitoring System Evaluation Report for the Pleasants Power Station* (Tetra Tech, 2017). For the LOEs in Checklist 2, the following evaluation criteria were used:

• Primary Indicators – As per Table A-1 in EPRI (2017), primary indicator constituents for CCRs include the CCR Rule parameters Boron (Appendix III), Calcium (Appendix III), Chloride (Appendix III), Fluoride (Appendix III and IV), Lithium (Appendix IV), Molybdenum (Appendix IV), and Sulfate (Appendix III), as well as Bromide, Potassium, and Sodium, which are parameters that are not listed in the CCR Rule.



- Secondary Indicators For this ASD, secondary indicator constituents for CCRs include those Appendix III and IV constituents that are not considered primary indicators.
- Leachate Data Analytical results from five leachate sampling events performed at the CCR unit between October 2017 and July 2019 at three locations (LM1, LM5, and LM7) were used for comparison to the February 2019 AM-3 groundwater results, as shown in Table 4. The comparison of data for barium and radium indicates that barium is found at higher concentrations in groundwater in both the upgradient well and in all the downgradient wells than in leachate, whereas radium is found at higher concentrations in only the downgradient wells than in leachate, indicating a localized, non-CCR source exists along the northern boundary of the CCR unit. Alternatively, concentrations of arsenic and lithium in the leachate samples are several times higher than those of the upgradient well and the downgradient wells, indicating that the arsenic and lithium SSLs in groundwater are likely attributable to a release from the CCR unit.
- Site Hydrogeology As discussed in the *CCR Rule Groundwater Monitoring System Evaluation Report* (Tetra Tech, 2017), groundwater in the CCBDF area occurs primarily within the fractured bedrock of multiple Conemaugh Group sandstone units including the Morgantown, Grafton, Jane Lew, and Saltsburg, which have been collectively identified as the uppermost aquifer for CCR Rule groundwater monitoring for the combined landfill and impoundment units. The CCR groundwater monitoring well network at the site is shown on Figure 1 and consists of three upgradient (background) wells (GW-7, -21, and -22), six downgradient wells to monitor the northern side of the combined CCR units (GW-19, -20, -23, -24, -25, and -26), and four downgradient wells to monitor the western side of the combined CCR units (GW-9, -27, -28, and -29). Historic and recent groundwater level data indicate groundwater flow at the site as flowing north from the topographically higher areas located to the south and southeast of the impoundment. Groundwater flow northwest of the dam and under the landfill is in the downstream direction of McElroy's Run toward the west. Flow in all of the rock units exhibit little seasonal and temporal fluctuations.

Having sufficient recoverable volumes of groundwater from one of the upgradient (GW-21) and three of the downgradient wells (GW-23, -24, and -25) was found to be problematic during both the background and initial DM sampling events. These four wells were noted to have low to very low yields during their installation and development which was anticipated given that historical well borings drilled at the site under the WVDEP



groundwater monitoring program were abandoned over time due to a lack of water in the same rock units. During the initial DM sampling event, sufficient recoverable groundwater volumes were found to be available in GW-23 and -24 but not in GW-21, -25, or in an additional downgradient well, GW-26. Geologic and hydrogeologic characteristics of the site, the monitoring well network, and the initial DM results are discussed in greater detail in both Tetra Tech 2017 and 2018.

It was originally intended that upgradient wells GW-21 and GW-22, which are both screened in the Morgantown sandstone, would be grouped for statistical evaluation purposes. However, after both the background and the initial DM sampling events were completed, it was determined that the two wells did not have the level of statistical similarity needed for grouping and that the availability of sufficient volumes of recoverable water was a recurring problem for GW-21. As such, it was decided that only GW-22 would be used to establish background chemistry for the northern side of the CCR units since it exhibited lower concentrations of all the Appendix III parameters than those measured in GW-21 and it also provided a reliable water yield while GW-21 did not. GW-21 was left in place (i.e., it was not abandoned) and it has been sampled when sufficient volumes of recoverable water were available. GW-21's water levels have also continued to be used to verify groundwater flow patterns at the site. FE intends is to keep GW-21 as a part of the CCR monitoring network until a sufficiently-sized data set can be compiled and used to determine whether or not it's statistically appropriate to group its results with the data set for GW-22. As discussed in Section 1.0, recent groundwater elevation measurements and mapping of the potentiometric surface indicate that GW-7, instead of a combination of GW-7 and GW-22 for the western and northern boundaries, respectively, acts as the upgradient well for the CCR network for both the western and northern boundary CCR wells as shown on Figure 2.

• CCR Unit Design - As shown on Figure 1, the CCR unit consists of two conterminous disposal areas, an impoundment and a landfill, that share a common boundary (the impoundment dam). The majority of the CCR material that has been disposed of at the site is managed in an unlined impoundment formed by a dam constructed across McElroy's Run. The dam was constructed with a clay-filled cutoff trench at the upstream toe and a clay blanket on the upstream face to function as a low permeability barrier. The downstream portion of the dam was constructed using compacted fly ash and periodic layers of bottom ash for blanket drains connected to sloping chimney drains that collect



seepage to discharge pipes for monitoring. The downstream face of the dam is covered by the landfill facility which WVDEP considers to be a buttress to the dam.

The landfill consists of three primary development stages which are further subdivided into construction subareas. At this time, development and disposal operations have only been performed in Stages 1 and 2 and the Stage 3 area remains undeveloped. Up until 2009 all of the landfill subareas were constructed with a compacted clay liner system that included an underlying combined groundwater underdrain/leak detection system and an overlying leachate collection system. Since 2009 a composite geosynthetic liner system (geosynthetic clay liner and geomembrane) has been utilized which also includes an underlying combined groundwater underdrain/leak detection system and an overlying leachate collection system. For all portions of the landfill that overlie the downstream face of the impoundment dam, a bottom ash blanket drain layer has also been utilized. Leachate and contact stormwater runoff from the landfill disposal areas are managed in Sedimentation Pond Nos. 1 and 2, which are lined impoundments located immediately down-valley of the future Stage 3 landfill development area. These impoundments also accept flows from the groundwater underdrain/leak detection zones and stormwater runoff from portions of the landfill's South Haul Road. Discharges from Sedimentation Pond Nos. 1 and 2 are pumped up to the CCR disposal impoundment and, ultimately, routed through the impoundment's dewatering system.

Based on the various LOE findings presented in Table 2, arsenic and possibly lithium SSLs determined for the AM-1, -2, and -3 events can most likely be attributed to a release from the CCR unit. However, the comparison of leachate data to upgradient and downgradient wells indicates that a source other than the CCR unit may be contributing to the occurrence of barium and radium in groundwater.

## **3.3 ASD CHECKLIST 3**

ASD Checklist 3 is attached as Table 3 of this report. The checklist evaluations were performed similar to those of ASD Checklist 2 by re-reviewing the groundwater analytical results (background, DM, and AM) for both Appendix III and IV parameters, leachate data for the CCR unit (specifically for barium, lithium, and radium) provided by FE, and hydrogeologic and design information and data included in *CCR Rule Groundwater Monitoring System Evaluation Report for The Pleasants Power Station* (Tetra Tech, 2017). For the LOEs in Checklist 3, the following evaluation criteria were used in addition to those used for ASD Checklist 2:

#### CCR Rule Appendix IV ASD Report **CCR Rule Appendix IV ASD Report** CCR CCR Rule Appendix IV ASD Report October 2019 2018/2019 Assessment Monitoring – Pleasants

- Results of AM/Nature and Extent of Release (N&E) groundwater sampling conducted in February and July 2019 indicate that an alternate source of barium, lithium, and radium appears to exist along the northern boundary as shown on Figures 3, 4, and 5, respectively. Isoconcentration contour lines located around these northern boundary wells indicate a localized source of all three parameters in this area. Historical and current oil and gas exploration and production activities have occurred in this area and are documented sources of barium, radium, and lithium that could be the source of the SSLs in the northern boundary wells. These results and associated comparisons are discussed in greater detail in Section 3.5 of this report.
- Review of site-wide boring logs for observations of potential oil and gas well impacts to groundwater during previous investigations identified several wells in which oil and gas impacts were noted. Observations of petroleum/hydrocarbon odor, sheen, and/or crude oil product were noted for the following wells at the time of their installation (copies of the relevant pages from each log are included as Attachment A of this report):
	- GW-3 light hydrocarbon odor
	- $-GW-4 oil$  odor
	- GW-5 oil odor and sheen
	- GW-6 black crude in rock cuttings
	- GW-7 hydrocarbon odor, black crude in rock cuttings
	- $P-96-4 oil$  odor
	- $P-96-5$  crude oil odor
	- $N-3$  oil odor
	- GW-13 crude oil in sandstone, visual staining
	- GW-15 0.32 feet of crude oil-fingerprinted product
	- GW-19 crude oil odor
	- GW-24 petroleum hydrocarbon odor
	- GW-25 petroleum hydrocarbon odor

Based on the LOE findings presented in Table 3 and the discussion above, the barium, radium, and lithium SSLs determined for the AM-1, -2, and -3 events can most likely be attributed to historical and current oil and gas exploration and production activities. While lithium has also



been shown to be a component of oil and gas well brine, the relatively high concentrations of lithium in the leachate is an indication that the CCR unit may be the source of the lithium SSLs.

### **3.4 REGIONAL GROUNDWATER STUDY**

In an effort to evaluate the natural variation in groundwater quality in the various water producing units of the Conemaugh Group (e.g., Morgantown, Grafton, Jane Lew, and Saltsburg sandstones) which comprise the CCR Rule uppermost aquifer, *Ground-Water Hydrology of the Minor Tributary Basins of the Ohio River, West Virginia* (USGS, 1984) was reviewed. The report review did not yield any specific information regarding natural variation of arsenic, barium, lithium, or radium in regional groundwater. However, the following table presents the range and mean concentrations reported for Appendix III constituents with SSIs in the Conemaugh Group wells which can be compared with CCR unit well data that point to oil and gas exploration activities as an alternative source:



Based on these reported values, the following observations were made:

- **Chloride -** The reported mean concentration of 31 mg/L is below the UPL for upgradient well GW-7 (104 mg/L), and the reported maximum concentration of 130 mg/L is slightly higher than the GW-7 UPL. With respect to downgradient wells along the northern boundary with Appendix IV SSLs, the reported maximum chloride concentration of 130 mg/L is well below the concentrations of chloride in GW-23 (12,900 mg/L), GW-24 (8,520 mg/L), and GW-25 (7,110 mg/L).
- **Sulfate** Sulfate concentrations tend to have an inverse relationship with other parameters typically present in groundwater impacted by oil and gas activities. Accordingly, the reported minimum concentration of 10 mg/L is significantly higher than both the GW-7 UPL of 0.5 mg/L and the sulfate concentrations in downgradient wellsGW-23 (0.2664 mg/L), GW-24 (<0.0386 mg/L), and GW-25 (0.618 mg/L).
- **TDS** The reported mean concentration of 371 mg/L is well below the UPL for GW-7 (1,260 mg/L). The reported maximum TDS concentration of 589 mg/L is also well below



the GW-7 UPL. With respect to downgradient wells with Appendix IV SSLs, the reported maximum TDS concentration of 589 mg/L is well below the concentrations of TDS for GW-23 (68,500 mg/L), GW-24 (42,400 mg/L), and GW-25 (35,900).

The comparisons noted above indicate that upgradient chloride and TDS concentrations (all indicators of oil and gas brine) at the site appear to be higher than the concentrations measured in regional Conemaugh Group groundwater during the USGS study period, while upgradient sulfate concentrations appear to be within the range of or below the concentrations measured in the study. However, comparing the maximum reported study results to the results for the corresponding downgradient wells with Appendix IV SSL concentrations indicates that all of the wells exhibit chloride and TDS concentrations that are higher to much higher than those for regional groundwater. Reduced sulfate, elevated chloride and, to a lesser extent, elevated TDS concentrations are typically observed with oil and gas exploration and production activities as discussed in the following section.

## **3.5 POTENTIAL FOR OIL AND GAS WELL IMPACTS**

In an effort to evaluate the potential for oil and gas well development on and near the site to have impacted groundwater for the SSL constituents, particularly barium, lithium, and radium, and to substantiate the results of Checklist 3, several lines of evidence related to oil and gas impacts were evaluated including a review of nearby oil and gas wells and their completion records, historical research related to oil and gas exploration activities near the site, research related to the occurrence of the site's SSL constituents in oil and gas activities, and historical investigations and studies performed at the site regarding oil and gas impacts.

### **3.5.1 Nearby Oil and Gas Well Locations and Completion Information**

The locations of oil and gas wells and basic information on the wells (e.g., total depth, date drilled, status, etc.) were obtained from the West Virginia Geologic and Economic Survey (WVGES) online oil and gas well database (http://ims.wvgs.wvnet.edu/WVOG/viewer.htm). Figure 6 presents the locations of these wells relative to the CCR monitoring well network and includes field observations of existing on-site oil and gas wells and associated infrastructure as well as groundwater sampling field notes that indicate oil and gas well-related impacts (e.g., sheen, odor, free product). A total of more than 100 existing or plugged/abandoned oil and gas wells were identified as shown on Figure 6. The table below summarizes key information for these wells obtained from the online database records:



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Note: Wells having API #s from 4707390041 through 4707390140 are also listed but have no associated information.

The completion dates for most of the wells are unknown, implying they were drilled as part of historic oil and gas well exploration in the area and potentially could have been drilled in the early 1900s or possibly in the late 1800s. A review of data for the other wells indicates they were drilled between 1935 and 2011. The total depths of the wells range from 71 ft to 5,684 ft and they've produced from formations including undifferentiated Upper Devonian Sandstone units. Many of the wells are reported as orphan wells and some have little or no information provided. As indicated on Figure 6, the wells are distributed across much of the site and adjoining areas. Considering the age of the wells there would seem to be potential for groundwater impacts from corroded/damaged well casing, degraded seals, etc., which could result in out-of-interval migration of oil and gas and formation brine. Any leaking oil and gas gathering lines/pipelines and wellhead brine storage tanks at currently producing locations could be another potential



source of releases. As discussed further below, potential constituents known to be associated with oil and gas wells include barium, radium, chloride, sodium, lithium, and elevated TDS levels.

### **3.5.2 Occurrence of SSL Constituents in Oil and Gas Brines**

It is noted in the "Chemistry and Origin of Oil and Gas Well Brines in Western Pennsylvania," (Dresel, P.E., and Rose, A.W., 2010) that brine samples collected from oil and gas operations indicate "…radium shows a general correlation with barium and strontium and an inverse correlation with sulfate." The data presented in Section 3.4, in which sulfate concentrations are inversely low compared to barium concentrations, supports this conclusion. The following table presents the range and mean concentrations reported in Dresel and Rose (2010) for applicable Appendix III/IV constituents in western Pennsylvania brines (assumed to be similar to those in West Virginia based on age and depositional environment):



Based on these reported values, the following observations were made:

- **Barium -** The reported mean concentration of 877.37 mg/L is well above the UPL for upgradient well GW-7 (0.0934 mg/L). With respect to downgradient wells with SSLs for barium, the reported mean concentration of 877.37 mg/L is also well above the concentrations of barium in GW-23 (9.76212 mg/L), GW-24 (9.25331 mg/L), and GW-25 (7.62675 mg/L). However, brine impacts to those wells would be expected to be diluted by groundwater and, hence, a potential reason they are lower.
- **Chloride -** The reported mean concentration of 104,544 mg/L is three orders of magnitude greater than the UPL for upgradient well GW-7 (104 mg/L), and the reported minimum concentration of 5,760 mg/L is also higher than the GW-7 UPL. With respect to downgradient wells along the northern boundary with Appendix IV SSLs, the reported minimum chloride concentration in brines of 5,760 mg/L is below the concentrations of chloride in GW-23 (12,900 mg/L), GW-24 (8,520 mg/L), and GW-25 (7,110 mg/L)

indicating the groundwater in those wells is within the range of the minimum and maximum concentrations of chloride found in brine.

- **Lithium** The reported mean concentration of 61 mg/L is significantly higher than the GW-7 UPL of 0.023374 mg/L. With respect to the downgradient well with an SSL for lithium, the reported mean concentration of 61 mg/L is higher than the concentration of lithium in GW-23 (0.150178 mg/L). However, brine impacts to GW-23 would also be expected to be diluted by groundwater and, hence, a potential reason they are lower.
- **Radium 226** The reported mean concentration of 2,150 pCi/L is significantly higher than the GW-7 UPL of 0.58 pCi/L for the sum of both radium-226 and radium-228. With respect to downgradient wells with Appendix IV SSLs, the reported mean radium-226 concentration of 2,150 pCi/L in brine is higher than the concentration of radium-226 in GW-23 (23.6 pCi/L), GW-24 (12.7 pCi/L), and GW-25 (13.2 pCi/L). However, brine impacts to GW-23, GW-24, and GW-25 would also be expected to be diluted by groundwater and, hence, a potential reason they are lower.

An additional study regarding the occurrence of radium with oil and gas produced waters conducted by the USGS identified median radium concentrations of 2,460 pCi/L and 734 pCi/L, for Marcellus Shale and non-Marcellus Shale produced water samples, respectively (USGS, 2011). An increase in concentration of radium was directly correlated with increases in TDS and salinity of the produced water.

## **3.5.3 Previous Oil and Gas Impact Studies at the Site**

In March 2004, Hydrosystems Management, Inc. (HMI) prepared a report for Allegheny Power Supply Company (a predecessor company of FirstEnergy) which evaluated increased barium concentrations in groundwater samples from monitoring well GW-4. GW-4 is part of the state Solid Waste/NPDES groundwater monitoring system, is located in the north-northeastern portion of the site (as shown on Figure 1), and has a total depth of 255 feet and a screen length of 55 feet. Barium concentrations in the well consistently exceeded the Ground-Water Quality Standard (GWQS) established in the facility's Solid Waste/NPDES permit. The HMI report concluded that leakage of brine from surrounding oil and gas wells was the most probable cause of the barium GWQS exceedances. GW-4 also showed increases in sodium and chloride levels. The HMI report indicated two known oil and gas wells were within 1,000 feet of GW-4 and referenced the existence of numerous orphaned wells in the area. As noted in Section 3.3 of this report, the boring log for GW-4 indicated oil and gas odors at the time of drilling; additionally, some oil



associated with groundwater and oil sheen were both present during well installation and development.

In 2017, oil observed in GW-23 sample water was submitted for fingerprinting laboratory analysis to determine the exact oil type. Results of that fingerprinting analysis indicated that the oil from GW-23 was representative of a straight chain hydrocarbon mineral oil. This oil is likely a result of historical oil and gas exploration activities that have occurred in the area over the past 150 years. A copy of the fingerprinting analysis results is provided as Attachment B.

### **3.5.4 Historical Oil and Gas Activities in the Surrounding Area**

Historical references regarding local oil and gas exploration activities in the Pleasants County area were also reviewed. In "A History of Pleasants County, West Virginia," (Pemberton, 1929) the Burning Springs-Eureka anticline is noted as having its "ridge" eroded and exposing lower (older) strata with oil-bearing rocks located at or near the surface. Additionally, the First Cow Run sand mentioned in the text (from which oil and gas have been produced) is also known as the Saltsburg Sandstone, the formation in which numerous on-site wells have penetrated. Bearing more relevance to the site is the following anecdote:

"Brown and Company of New York drilled in a well on McElroy Run back of Eureka on the Giles Hammett farm, which came to be known as the 'Burnt Well,' heretofore mentioned. At a depth of 1,100 feet a copious quantity of oil was found filling the hole to a depth of 100 feet. This was on April 27, 1886. A few days later the well was shot, and for a time flowed at a rate of forty barrels a day. Unfortunately, the rig caught fire, the cable was burned, and the heavy tools fell into the hole, where they remained about a year."

The 1974 Environmental Impact Statement (EIS) (U.S. Army Engineer District, 1974) completed for the Pleasants Power Station noted that several oil and gas wells were drilled in 1958 and 1959 in the vicinity of the plant with one drilled to 740 feet producing 11 barrels of oil the first day. Four additional wells drilled to depths between 1,600 and 2,527 feet produced similar quantities of oil. It was stated in the EIS that "…it is presumed locally that these oil wells are those which have contaminated the water wells in the site area."

In summary, the potential for impacts to groundwater by oil and gas wells on the site and in nearby upgradient areas appears to be significant, particularly in light of the historical and welldocumented oil and gas well impacts in many of the groundwater monitoring wells located onsite. The data presented in this section indicate that the Appendix IV parameters barium and radium are likely attributable to oil and gas (brine) impacts. Lithium, which was reported at very



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high concentrations in oil and gas well brines for formations present at the site, may also be related to oil and gas brines, but since it is also present in site leachate at concentrations well above concentrations reported in the upgradient and downgradient CCR monitoring wells, it is not possible to clearly differentiate the source of lithium SSLs. However, as indicated by comparing the radium and barium isoconcentration maps (Figures 3 and 4, respectively) with the lithium isoconcentration map (Figure 5), the location of the highest concentrations for all three of these constituents occurs at GW-23, located along the northern property boundary, suggesting that lithium may exhibit a potential relationship with the barium and radium impacts from oil and gas well activities. Additionally, wells immediately downgradient of the leachate collection system along the western boundary (GW-27, GW-28, and GW-29) do not exhibit elevated concentrations of lithium, barium or radium, indicating that the presence of the three constituents in concentrations greater than their respective GWPS along the northern boundary are likely correlated and associated with oil and gas well impacts.



### **4.0 CERTIFICATION STATEMENT**

In accordance with § 257.95(g)(3)(ii) of the CCR Rule, an ASD for Appendix IV constituents was undertaken for the CCR unit identified herein. Based on the information and data that were available for review, the following determinations have been made with respect to the AM-1, -2, and -3 events:

- The barium and radium SSLs can be attributed to historical and current oil and gas exploration and production activities that have occurred on-site. As such, in accordance with the applicable requirements of § 257.95 of the CCR rule, no corrective measures are required for these parameters and assessment monitoring for barium and radium will continue.
- The lithium SSLs are currently considered indeterminate based on the LOE's presented herein, but the available evidence indicates a high potential for the elevated lithium concentrations to also be attributable to oil and gas impacts at the site based on the occurrence of the barium, radium, and lithium concentrations above the GWPS occurring in the northern boundary in which extensive oil and gas activities have occurred historically. To resolve this uncertainty, the applicability of leachate and groundwater lithium isotopic analysis at the site will be evaluated and lithium sampling of brine from onsite production equipment will be considered. Pending completion of that work and for the purposes of this ASD, lithium has not been categorized as attributable to either the CCR unit or to an alternate source. It will continue to be analyzed as part of the assessment monitoring program and will transition to the applicable requirements of assessment of corrective measures per § 257.96 of the CCR Rule, should isotopic analysis and/or brine sampling indicate the CCR unit is the likely source of the lithium exceedances.
- The arsenic SSLs could not be attributed to sources other than the CCR unit, to errors in sampling, analysis, or statistical evaluation, or from natural variation in groundwater quality. As such, a transition to the applicable requirements of assessment of corrective measures for arsenic per § 257.96 of the CCR Rule appears to be warranted and assessment monitoring of this parameter will also continue.



### **5.0 REFERENCES**

- Dresel, P.E., and Rose, A.W., 2010. *Chemistry and Origin of Oil and Gas Well Brines in Western Pennsylvania.* Pennsylvania Bureau of Topographic and Geologic Survey, 4<sup>th</sup> ser., Open-File Report OFOG10-01.0, 48 p., Portable Document Format (PDF).
- EPRI, 2017. *Guidelines for Development of Alternative Source Demonstrations at Coal Combustion Residual Sites*. EPRI, Palo Alto, CA: 2017. 3002010920.
- HMI, 2004. *Evaluation of Increased Barium Concentrations in Groundwater Samples From the GW-4 Monitoring Well at the McElroy's Run Disposal Facility.* HMI, March 23, 2004.
- Pemberton, 1929. *A History of Pleasants County, West Virginia.* Robert Pemberton, St. Mary's, WV, 1929. Reprinted for Clearfield Company, Inc. by Genealogical Publishing Co., Inc., Baltimore, MD, 2002.
- Tetra Tech, 2017. *Coal Combustion Residuals (CCR) Rule Groundwater Monitoring System Evaluation Report, Pleasants Power Station, Coal Combustion Byproduct Landfill and Impoundment*. Tetra Tech, Inc., Pittsburgh, PA, October 2017.
- Tetra Tech, 2018. *2017 Annual CCR Groundwater Monitoring and Corrective Action Report, McElroy's Run Coal Combustion Byproduct Disposal Facility, Pleasants Power Station*. Tetra Tech, Inc., Pittsburgh, PA, January 2018. http://ccrdocs.firstenergycorp.com/
- Tetra Tech, 2019. *2018 Annual CCR Groundwater Monitoring and Corrective Action Report, McElroy's Run Coal Combustion Byproduct Disposal Facility, Pleasants Power Station*. Tetra Tech, Inc., Pittsburgh, PA, January 2019. http://ccrdocs.firstenergycorp.com/
- U.S. Army Engineer District, 1974. *Draft Environmental Impact Statement Pleasants Power Station Units No. 1 and 2*. U.S. Army Engineer District, Huntington, WV, June 1974.
- USGS, 1984. *Ward, P.E., and Wilmoth, B.M., Ground-Water Hydrology of the Minor Tributary Basins of the Ohio River, West Virginia. West Virginia Geological and Economic Survey Basic Data Report 1.*
- USGS, 2011. Rowan, E.L., Engle, M.A. Kirby, C.S., and Kraemer, T.F.*, Radium Content of Oiland Gas-Field Produced Waters in the Northern Appalachian Basin (USA): Summary and Discussion of Data, Scientific Investigations Report 2011-5135,* U.S. Department of the Interior*.*

# **TABLES**







Field sheeps identified sheep in GW-25 for Events 4 through 13, and GW-25 for Events 4

-24, -25, and -26 due to limited available water.

V-22 (Events 1 and 8 through 13), GW-24 (Event

and 4, GW-25 for Events 1 and 7 through 10, and 4

lank so no action taken. Arsenic detected in lab 27, -28, and -29 qualified "U". In Event 11, Ra228

Barium carrier gas had radiation counts > limit, so qualified "J".

d high recovery for MS/MSD for As in Event 4 duplicate, GW-28, and -29) qualified "J" due to

to statistically prove a false positive result without

6/228. In downgradient wells used for AM-1, AM-2<br>, and -29.

### **Table 2 - ASD Checklist 2: Lines of Evidence Associated with the CCR Unit**

and -24; No Sulfate SSIs. Is in GW-9, -27, and -29.

and  $-29$ .

, Lithium, and Molybdenum are all found at oring wells.

Lithium, and Molybdenum are all found at oring wells.

Chloride (GW-19, -20, -23, and -24), Fluoride (GWve exhibited SSIs. Lithium is an SSI in GW-24 and

 $-29$ ) and Chloride (GW-27,  $-28$ , and  $-29$ ) have exhibited SSIs. Lithium has exhibited SSIs in GW-29; Molybdenum has exhibited SSIs in (GW-28).

> ide – Yes; Fluoride - No. It is noted that statistical sults: evaluation is based on four leachate  $017$  and April 2019.

> fate – Yes. It is noted that statistical analysis has ion is based on four leachate sampling events

the CCR dataset covers a very limited time range

the CCR dataset covers a very limited time range

4), TDS (GW-19, -20, -23, and -24), Barium (GW-19 and GW-20), Chromium (GW-20), Radium 226+228 (GW-9 and -19), and Selenium (GW-20); SSLs for Arsenic (GW-19, -23, -24, and -25), Barium (GW-23, -24, and -25), and Radium 226+228 (GW-9





nd -29), TDS (GW-28 and -29), Barium (GW-27, -, and -29); SSLs for Arsenic (GW-29).

s; Barium – No; Radium 226+228 not historically pled once in July 2019 for this ASD. Statistical

s; Barium – No; Radium 226+228 not historically pled once in July 2019 for this ASD. Statistical

the CCR dataset covers a very limited time range

the CCR dataset covers a very limited time range

sting program at site.

listed above, major chemistry analysis was not

listed above, isotopic analysis was not performed

wngradient wells, all of which are positioned dof the year.



ernative Source (Row b)

ative Source (Row b)

parison omparison .<br>Sow a

ernative Source (Row b)

ernative Source (Row c) omparison kow a)

olluvial soils, mostly along what were the

the waste boundary except for GW-23 (Northern


Key lines of evidence are based on relationships that must be observed in order for an SSI/SSL to be due to a release from a CCR unit. If these relationships are not observed, then they are critical to establishing an ASD. It is difficult to build a strong ASD without any key lines of evidence. It may be possible to build an ASD with a single key line of evidence, but the ASD will be stronger with additional key or supporting lines of evidence.

Supporting lines of evidence provide additional information that supports the ASD. Supporting lines of evidence are generally not sufficient to build an ASD unless there is at least one key line of evidence, although it may be possible if there are many supporting lines of evidence.

<sup>3</sup> This LOE applies to:

Table Notes:

<sup>1</sup> ND (not determined) indicates that this line of evidence was not tested or there are insufficient data to make a determination; N/A means lines of evidence not applicable to the CCR unit.

<sup>2</sup> Line of Evidence (LOE) Types:

Constituent: An SSI/SSL for that constituent at any monitoring point

Monitoring Point: All SSIs/SSLs at a specific monitoring point

Case: An SSI/SSL for a specific constituent at a specific monitoring point

Unit: All SSIs/SSLs at the monitored unit

site is comprised of multiple water-bearing strata adient well (GW-7) is located along the appropriate gradient wells, however, it is are also positioned adient wells.

the impoundment area (including the dam) does

inches of compacted clay, while the remainder ynthetic clay liner (GCL) overlain by a high density

a leachate collection system. The impoundment but the dam does include a blanket drain/chimney

ated within a valley (the impoundment at the head defined uppermost aquifer at the site is comprised cally connected. Most of the uppermost aquifer rock strata all outcropped within the valley before the disposal site was developed so it is very likely Iy in the impoundment area.

### **Table 3 - ASD Checklist 3: Lines of Evidence Associated with Alternative Natural and Anthropogenic Sources**

### **Indicative** Determination / Basis

and gas exploration and production activities have ne water and associated constituents of concern tred aquifer. Several hundred oil and gas wells dating back as far back as the late 1880s have the potential to have been d, or produced, resulting in releases to the

> and gas exploration and production activities have  $\epsilon$  ounding the CCR unit, including areas f the monitoring locations.

onitoring locations situated between the potential he CCR unit.

al pile runoff, or coal pile leachate management lient monitoring points.

mining operations that have occurred on-site or in

units located upgradient or side gradient of the ions.

refer to Table 2, LOE 5b) and constructed atop the downstream face of the unlined impoundment's dam. However, the two disposal areas share a multi-unit groundwater monitoring network that ntiation of impacts from one area or the other.



### **Indication LOE Type<sup>2</sup> Applies to<sup>3</sup> Weight of Evidence Determination / Basis**

landfill area does have a combined groundwater underdrain/leak detection system and the impoundment dam has a blanket em. However, the impoundment area does not have any type of groundwater control system. As such, the landfill system i measurable localized influence on groundwater

> $\frac{1}{2}$  dust suppression water to have flowed off the iner or runoff containment systems and near

ined from non-potable sources from the power

27 and -28 are located near the CCR landfill haul

-27, -28, and -29 are positioned downgradient of sluice line and effluent siphon line.

27 is located near the landfill's leachate collection ines.

ills of CCRs, leachate, or sluice water in cations.



of the impoundment dam is constructed of compacted fly ash and includes blanket and chimney drains that are

### ted downgradient and distant from the CCR

opermost aquifer at the site is comprised of multiple water-bearing strata that are hydraulically connected. The site's upgradient well (GW-7) and other background wells (GW-21 and -22) are located along the appropriate groundwater flow paths to the downgradient also positioned stratigraphically higher than s wells.

of cyclic sequences of sandstone, shale, claystone, coal, and limestone and is not considered to be poorly buffered.



### *<u>I* Evidence Determination / Basis</u>

nent activities at the site have been performed in nt disposal areas..

orical and current oil and gas tank batteries and underground pipelines on the site with at least one known release from an near GW-7 approximately 15 years ago.

> cess road that is paved and salted is located downit monitoring wells.

by droseeding of all disturbed areas at the site eas, etc.)

### **Indication LOE Type<sup>2</sup> Applies to<sup>3</sup> Weight of Evidence Determination / Basis**

luenced by pH and redox; sorption usually

ited solubility and is usually sorbed to clay, soils,

t to cation exchange.

not analyzed as part of the Appendix IV ASD.

have generally remained consistent with changes not being attributable to flooding and drought conditions.

isistently been to the north and west and to the and northern boundaries, respectively.

mine, mine spoil, or conveyor systems upgradient

ies appear to have limited agricultural uses mined to present little to no impacts to s to the CCR unit.

ies appear to have limited agricultural uses mined to present little to no impacts to s to the CCR unit.



### **Indication LOE Type<sup>2</sup> Applies to<sup>3</sup> Weight of Evidence Determination / Basis**

orical and current oil and gas production tank batteries surrounding the CCR unit. Documented spills from those tanks were not identified, but given the age of the tanks there is the potential that alted in impacts to groundwater.

d historical and existing oil and gas exploration and production wells on and in the vicinity of the site. Observations of oil and gas impacts to groundwater have been noted during the installation of itoring wells at the site and during groundwater

> s activities discussed in LOE 9f, there are no site commercial and/or industrial sources.

and gas exploration and production activities have ration of brine water and other constituents of interest in the overlying aquifer of the CCR unit that could be affecting

> R unit and history of disposal activities dating back as been enough time for potentially impacted e affected monitoring wells.



Supporting lines of evidence provide additional information that supports the ASD. Supporting lines of evidence are generally not sufficient to build an ASD unless there is at least one key line of evidence, although it may be possible if there are many supporting lines of evidence.

<sup>3</sup> This LOE applies to:

Table Notes:

 $^1$  ND (not determined) indicates that this line of evidence was not tested or there are insufficient data to make a determination; N/A means line of evidence not applicable to the CCR unit.

<sup>2</sup> Line of Evidence ( $LOE$ ) Types:

 Key lines of evidence are based on relationships that must be observed in order for an SSI/SSL to be due to a release from a CCR unit. If these relationships are not observed, then they are critical to establishing an ASD. It is difficult to build a strong ASD without any key lines of evidence. It may be possible to build an ASD with a single key line of evidence, but the ASD will be stronger with additional key or supporting lines of evidence.

Constituent: An SSI/SSL for that constituent at any monitoring point

Monitoring Point: All SSIs/SSLs at a specific monitoring point

Case: An SSI/SSL for a specific constituent at a specific monitoring point

Unit: All SSIs/SSLs at the monitored unit

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### **Table 4 - Leachate Data Summary**





Notes: DG -Downgradient; GW - Groundwater; UG - Upgradient; UPL - Upper Prediction Limit

Leachate Concentrations averaged from 5 sampling events performed between October 2017 and July 2019, except for Lithium and Radium which was from one event in July 2019.

GW Concentrations of App. III parameters from sampling and analysis completed in February 2019.

GW Concentrations of App. IV parameters from sampling and analysis completed in February 2019.

UG UPL's based on 8 baseline sampling events.

LM1 - Leachate Collection from Dam Blanket/Chimney Drains

LM5 - Stage 1G LCS

LM7 - Stage 2B LCS

## **FIGURES**



PGH P:\GIS\FIRST\_ENERGY\MAPDOCS\PLEASANTS\_POWERSTATION\_\_BORING\_PIEZOS\_WELLS\_CCR\_20190801.MXD 10/10/19 PD



PGH P:\GIS\FIRST\_ENERGY\MAPDOCS\PLEASANTS\_POWERSTATION\_PROPOSED\_JULY2019\_CCR.MXD 08/14/19\_PD



PGH P:\GIS\FIRST ENERGY\MAPDOCS\PLEASANTS POWERSTATION BARIUM FEBRUARY2019.MXD 09/13/19 PD



PGH P:\GIS\FIRST ENERGY\MAPDOCS\PLEASANTS POWERSTATION RADIUM FEBRUARY2019.MXD 08/09/19 PD



PGH P:\GIS\FIRST ENERGY\MAPDOCS\PLEASANTS POWERSTATION LITHIUM FEBRUARY2019.MXD 09/13/19 PD



PGH P:\GIS\FIRST\_ENERGY\MAPDOCS\PLEASANTS\_OG\_MW\_LOCS\_20190618.MXD 09/13/19 PD



## **ATTACHMENT A**

**Boring Logs with Observations of Potential Oil and Gas Well Impacts** 







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 $\sim$ 

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<u> 1987 (1986)</u>



 $\sim 10^{-1}$ 







![](_page_92_Picture_15.jpeg)

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![](_page_93_Picture_21.jpeg)

HO ROCK CORING WITH WATER REMARKS\*\*

PROJECT NO. 81-237-66 BORING NO.  $P-96-4$ 

\*POCKET PENETROMETER READINGS

\*\* METHOD OF ADVANCING AND CLEANING BORING

![](_page_94_Picture_27.jpeg)

REMARKS\*\*

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**North** 

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\*POCKET PENETROMETER READINGS

\*\* METHOD OF ADVANCING AND CLEANING BORING

PROJECT NO. 81-237-72<br>BORING NO.  $P-96-5$ 

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REMARKS\*\*

\*POCKET PENETROMETER READINGS

PROJECT NO. 81-237-72<br>BORING NO. P-96-5

. \*\* METHOD OF ADVANCING AND CLEANING BORING

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GAI CONSULTANTS INC.

BORFHOLF NO  $N-3$ 

PAGE  $/7$  of  $/7$ 

 $\sim$ 

![](_page_97_Picture_2.jpeg)

cementation, and fracturing. In accordance with stress relief fracture theory, well yields are highest in the valleys, moderate on the hillsides, and minimal on the ridges (Shultz, 1984).

### **FIELD INVESTIGATION METHODS**

Seven new monitoring wells (GW-13, GW-14, and GW-16 through GW-20) were installed for this study in 1995 (Figure 3-1). The wells were installed at locations where the bedrock aquifer has the potential for significant fracture development due to stress relief. In addition, ten existing monitoring wells were sampled for the study, and numerous boring logs from previous studies were available for geologic interpretation.

Monitoring wells GW-13, GW-14, and GW-20 are located on the east side of the McElroy's Run watershed. The wells were aligned along an eastward-trending transect identified as a potential groundwater flow path from the impoundment toward the neighboring French Creek watershed. The location of the transect coincides with small tributary valleys in the two watersheds. Wells GW-13 and GW-20 were installed as a cluster in order to investigate vertical gradients and water quality near the impoundment. Well GW-14 is located about 600 ft farther along the transect from the impoundment than the cluster. Boring GW-15 was drilled about 500 ft farther along the transect than GW-14. However, a thin layer  $(0.34 \text{ ft})$  of floating petroleum, analyzed as crude oil, was encountered in the borehole, and the borehole was abandoned.

Wells GW-16, GW-17, and GW-18 were installed in the valley bottom downstream from the impoundment dam. These three wells, along with existing wells MP-3 and MP-4, form a transect along the valley bottom from the dam to the Ohio River valley. Wells GW-16 and GW-17 were installed as a cluster to investigate vertical gradients and water quality near the toe of the dam. The MP-3/MP-4 well cluster is located approximately 1500 ft downgradient from the GW-16/17 cluster. MP-4 is installed in the shallow bedrock aquifer; MP-3 is an overburden well installed in the McElroy's Run valley alluvium. GW-18 is a bedrock well sited at the base of the McElroy's Run valley, near its junction with the Ohio River valley

Well GW-19 is located north of the impoundment. The well is aligned with pre-existing well GW-3 along a potential flow path through the ridge that separates the impoundment valley from the Ohio River.

Construction of these wells included coring, drilling, geophysical logging and packer testing. Each of these operations is summarized below. Additional detail is provided in Appendix A.

![](_page_99_Picture_4.jpeg)

 $\lambda$ 

![](_page_100_Figure_0.jpeg)

11-21-2017 Z:\000\_Primary Work\First Energy\M-Tech\GW-24\GW-24\_5.bor 11-21-2017 Z:\000\_Primary Work\First Energy\M-Tech\GW-24\GW-24\_5.bor

![](_page_101_Figure_0.jpeg)

11-21-2017 Z:\000\_Primary Work\First Energy\M-Tech\GW-25\.bor files\GW-25\_5.bor 11-21-2017 Z:\000\_Primary Work\First Energy\M-Tech\GW-25\.bor files\GW-25\_5.bor

### **ATTACHMENT B**

**GW-23 Oil Fingerprinting Laboratory Report**

![](_page_102_Picture_4.jpeg)

![](_page_103_Picture_0.jpeg)

[ **B***A* Laboratory

# **ISO 9001 Registered** *BETA* **Laboratory**

Chemical Analysis

6670 Beta Dr., Mayfield Village OH 44143 (440)-604-9832

![](_page_103_Picture_199.jpeg)

A water sample from the Pleasants Ground Water 23-CCR location was submitted for water analysis but when the container was opened an oil film was present on the water's surface. The oil was extracted off the water and analyzed using a FT Infrared Spectrometer.

### **Results:**

1) The oil was identified and a straight chain hydrocarbon oil (mineral oil).

#### **Discussion:**

The oil was extracted off the surface of the water using a dropper and the water was removed from the residue. The oil was then analyzed on the FT Infrared Spectrometer. ATTACHMENT 1 shows the results.

The FT Infrared Spectrometer was calibrated with Standard Reference Material (SRM)1921b, which is a matte finish polystyrene film certified by the National Institute of Standards and Technology (NIST). There was no Sample Analysis Request / Chain of Custody submitted for this analysis.

#### **Material Test Equipment**

Instrument Model: Perkin Elmer Frontier FT-IR Spectrometer, BETA 0755, Calibration Due: 5/4/17

Analysis Performed By  $\leq$   $\neq$   $\neq$ 

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ATTACHMENT 1: FTIR Spectrographic Analysis of the oil removed from the surface of the Pleasants GW-23-CCR water sample indicates the oil is a straight chain hydrocarbon mineral oil. Instrument: Perkin Elmer Frontier FT-IR Spectrometer, BETA 0755, Calibration due 5/4/17 Performed by J. Hirsch on 4/27/17

![](_page_104_Figure_1.jpeg)

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### **APPENDIX B**

**Geologic Cross-Sections** 

![](_page_105_Picture_5.jpeg)

![](_page_106_Figure_0.jpeg)

7.5 MINUTE SERIES, U.S.G.S. TOPOGRAPHIC QUADRANGLE, WILLOW ISLAND, WEST VIRGINIA 1957, PHOTOREVISED 1976

**BEC** 

 $\mathcal{V}(\mathbb{R})$ 

![](_page_106_Picture_2.jpeg)

AND MONITORING WELL LO MCELROY'S RUN EMBANKMENT AN

570 Beatty Rd. • Pittsburgh,<br>Monroeville, Pa. 15146<br>412-856-6400

**ALLEGHENY POWER MONONGAHELA POWE PLEASANTS POWER** PLEASANTS COUNT

### LEGEND

![](_page_106_Picture_7.jpeg)

**MONITORING WELL LOCATION** 

![](_page_106_Figure_9.jpeg)

APPROX. GRAFTON SANDSTONE

![](_page_106_Picture_11.jpeg)

APPROX. AMES LIMESTONE-**LOWER HARLEM COAL** 

APPROX. PITTSBURGH RED BEDS

![](_page_106_Picture_14.jpeg)

**SEDIMENTATION POND** 

![](_page_106_Picture_16.jpeg)

A LINE FOR GEOLOGIC CROSS-SECTION  $(FIG. 2)$ 

500 ELEVATION CONTOUR-TOP OF **AMES LIMESTONE** 

![](_page_106_Picture_95.jpeg)

![](_page_107_Figure_0.jpeg)